

Long term changes in the ecosystem in the northern South China Sea during 1976–2004

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Abstract. Physical and chemical oceanographic data were obtained by seasonal monitoring along a transect (Transect N) in the northern South China Sea (nSCS) during 1976–2004. Fluctuations of DIN (dissolved inorganic nitrogen), seawater temperature (SST and T_{av} – average temperature of the water column), N:P ratio and salinity (S_{av} and S_{200} – salinity at the 200 m layer) exhibited an increasing trend, while those of T_{200} , DO, P, Si, Si:N and SSS exhibited a decreasing trend. The annual rates of change in DIN, DO, T and S revealed pronounced changes, and the climate trend coefficients, which was defined as the correlation coefficient between the time series of an environmental parameter and the nature number (namely 1,2,3,... n), were 0.38 to 0.89 and significant ($p \leq 0.01$ to 0.05). Our results also showed that the ecosystem has obviously been influenced by the positive trends of both SST and DIN, and negative trends of both DO and P. For example, before 1997, DIN concentrations in the upper layer were very low and N:P ratios were less than half of the Redfield ratio of 16, indicating potential N limitation. However after 1997, all Si:P ratios were >22 and the $N_{av}:P_{av}$ was close to the Redfield ratio, indicating potential P limitation, and therefore N limitation has been reduced after 1997.

Ecological investigation shows that there have been some obvious responses of the ecosystems to the long-term environmental changes in the nSCS. Chlorophyll-*a* concentration, primary production, phytoplankton abundance, benthic biomass, cephalopod catch and demersal trawl catch have increased. But phosphorus depletion in upper layer may be related to the shift in the dominant species from diatoms to dinoflagellates and cyanophytes. The ecosystem response was

induced by not only anthropogenic activities, but also global climate change, e.g. ENSO. The effects of climate change on the nSCS were mainly through changes in the monsoon winds, and physical-biological oceanography coupling processes.

In this study physical-chemical parameters were systemic maintained, but the contemporaneous biological data were collected from various sources. Regional response to global climate change is clearly a complicated issue, which is far from well understood. This study was made an attempt to tackle this important issue. For the aim these data were valuable.

1 Introduction

The South China Sea (SCS) is the largest semi-enclosed marginal sea in Southeast Asia with an area of about 3.5×10^6 km². Our study area is the northern SCS (nSCS), bounded by the mainland of China on the north and northwest sides, Taiwan Strait on the northeast, Taiwan Island and Bashi Strait on the east side, and the Hainan Island on the west side. The nSCS is connected to the East China Sea through Taiwan Strait, and it is connected to the open ocean through Luzon Strait, where a deep sill (>2000 m) allows effective water exchange with the western Pacific. The topography of the area is characterized by the incline from the coast of mainland China towards the southeast, with a gradient from the coastal zone (<50 m), continental shelf (<200 m), the slope and open sea (>200 m), to the deep sea (>3000 m) (Fig. 1).

The runoff from 29 rivers, with different sized input into the nSCS with total drainage area of 5.5×10^5 km², and an annual fresh water discharge of 3.8×10^{11} m³ (Han et



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al., 1998). Among them, the Pearl River is the largest with a drainage area of $4.3 \times 10^5 \text{ km}^2$ and a discharge of $3.3 \times 10^{11} \text{ m}^3 \text{ y}^{-1}$ (Han et al., 1998). It carries a large quantity of suspended solids ($8.3 \times 10^7 \text{ t y}^{-1}$, Han et al., 1998) and dissolved nutrients ($\text{N}=8.6 \times 10^4 \text{ t y}^{-1}$; $\text{P}=1.2 \times 10^4 \text{ t y}^{-1}$; $\text{Si}=184.3 \times 10^4 \text{ t y}^{-1}$, before 1998, Wang and Peng, 1996; and $\text{N}=19.14 \times 10^4 \text{ t y}^{-1}$, $\text{P}=0.8 \times 10^4 \text{ t y}^{-1}$ after 1998, SOAC, 2000, 2001, 2002, 2003, 2004) into the nSCS. The Pearl River plume extends offshore to cover a large area of the nSCS (Yin et al., 2001). During the dry season in winter, the river plumes extend westward along the coast of Guangdong. Due to the strong northeast monsoon; during the flood season in summer, the river plume extends well into the nSCS, and its southeastward and southward tongue can reach up to $17^\circ 00' \text{ N}$, 112° E , about 5 degrees, in latitude, away from the river mouth (Cai et al., 2007; Xue et al., 2001a, b).

The meteorological forcing over the nSCS is dominated by the East Asian Monsoon (Sadler et al., 1985). The upper ocean circulation follows closely the alternating monsoons (Wyrtki, 1961). During winter northeast monsoon, along the northern boundary, the warm and saline Kuroshio Current water with oligotrophic properties intrudes through Luzon Strait and flows westward along the continental margin of China to become the deep-water mass of the nSCS (Nitani, 1972; Shaw, 1991). The coastal water of the East China Sea flows southwestward through Taiwan Strait into the nSCS (Fang et al., 1998; Xue et al., 2004). On the contrary, during the summer southwest monsoon, the Guangdong Coastal Current flows eastward along the southern coast of mainland China, which eventually flows into the East China Sea through Taiwan Strait. The southwesterly winds also induce Ekman transport toward offshore and coastal upwelling. The deep water upwells and mixes with the upper water to form the SCS intermediate water, which flows out of the nSCS into the northwestern Pacific Ocean through Luzon Strait (Gong et al., 1992).

In the nSCS, the thermocline occurs all the year round, and the interannual change in its strength is pronounced (Yuan and Deng, 1997a, b; Shi et al., 2001). Previous studies have examined on variations in seawater temperature and salinity distributions (Yang and Liu, 1998; Yuan and Deng, 1998), dissolved oxygen distribution (Lin and Han, 1998), pollution status along the coast of the nSCS (Li and Chen, 1998) and the fisheries environment in the nSCS (Jia et al., 2005). Furthermore, it has been found that due to the combined effects of monsoons, topography, shape of the coastal line and the inertial effects, mesoscale eddies (Zeng et al., 1989; Xu et al., 2001; Li et al., 2003; Chen et al., 2005). Recent studies revealed that the effects of coupling between physical – chemical – biological oceanographic processes on phytoplankton biomass and production are important for understanding the influence on the long-term environmental changes and the ecosystem dynamics of the SCS (Liu et al., 2002, 2007; Ning et al., 2004).

However, the long term changes in environmental conditions and the responses of the ecosystem in this region have not been well documented yet. The objective of this study was to analyze the 29 y time series of multidisciplinary observational data obtained during 1976–2004, aiming at understanding how the environment has changed and how the ecosystem and biological resources have responded to the environmental changes in the nSCS.

It must be pointed out that the data set we adopted in present analysis is large and from various sources (using data in this study were systemic maintained for the physical-chemical parameters, but for contemporaneous biological data were collected from various sources, and we have collected data from the same period of investigation as best as one can. Inevitably, there are some mismatches between the scales of physical and chemical parameters reflecting the processes of environmental change and the spatial and temporal dimensions of biological investigations, due to the less frequency for the latter). Although the quality of data might vary throughout the long period of observation, these data were valuable for the long term changes in environmental conditions and the responses of the ecosystem in this region. Using these data, we can still find some disciplines about the response of the ecosystem to the environmental change in the SCS. Regional response to global climate change is clearly a complicated issue, which is far from well understood. This study made an attempt to tackle this important issue. And in order to improve our understanding, additional long-term study is mandatory.

2 Data and methods

In this study, data were obtained from winter and summer monitoring along transect N (Fig. 1, an observation transect, including six stations, crossing the nSCS, from the northwestern to southeastern), maintained by the survey team of the State Oceanic Administration (SOA), China during 1976–2004. These data include physical [seawater temperature (T) and salinity (S)] and chemical parameters [dissolved oxygen (DO), phosphate ($\text{PO}_4\text{-P}$), silicate ($\text{SiO}_3\text{-Si}$), dissolved inorganic nitrogen (DIN, including $\text{NO}_3\text{-N}$, $\text{NO}_2\text{-N}$ and $\text{NH}_4\text{-N}$)]. The parameters of T, S and DO data collection started from 1976; and the nutrients data ($\text{PO}_4\text{-P}$, $\text{SiO}_3\text{-Si}$, $\text{NO}_3\text{-N}$, $\text{NO}_2\text{-N}$ and $\text{NH}_4\text{-N}$) collection started from 1989. Seawater samples were collected using Nansen bottles from the surface, 5, 10, 15, 20, 25, 30, 35, 50, 75, 100, 150 and 200 m for T and S, and at the surface, 10, 20, 30, 50, 75, 100, 150 and 200 m for biogenic element determination. Seawater temperature was measured by using a reversing thermometer attached to the Nansen bottle, and salinity was measured using induction salinometer, according to SOAC (1975) and NBTS (1991). Nutrients (nitrate, phosphate and silicate) were analyzed by standard spectrophotometric method, and dissolved oxygen (DO) was analyzed by the Winkler method

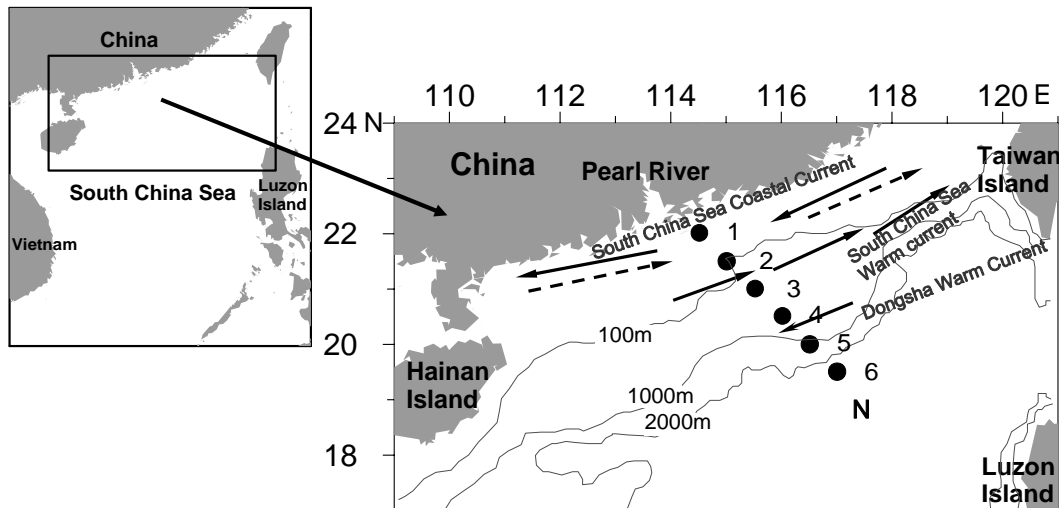


Fig. 1. Geographical locations of the transect and stations and the circulation in the northern South China Sea (nSCS) (modified from Su, 1998; Xue et al., 2004). Transect N from the Pearl River Estuary towards the southeastern nSCS is the main transect with 6 stations (full circle).

(Strickland and Parsons, 1972). Photosynthetic pigments (Chl-*a*) were measured by the acetone extraction and fluorescence method (Holm-Hansen et al., 1965).

Annual mean values were the average for winter and summer which were derived from observations during February and August, respectively. The regional average was the average value for all stations illustrated in Fig. 1. First, we took the values at the sea surface (SS), the depth of 200 m and the average through the water column for 0–200 m (integrated) for each parameter for each station, since at the depth of 200 m concentrations of biogenic elements and other properties were relatively stable and much less influenced by the upper layers. Second, the regional means for each parameter on an annual scale were calculated. The average value for the water column was computed, according to the following equation:

$$X_{av} = \frac{1}{b} \int_0^b X(z) dz \tag{1}$$

Where *X* is an environmental parameter; *b* is the water depth (200 m, or 2 m above bottom if the water depth is shallower than 200 m) and *z* is the observation depth.

In order to show the interannual changes in environmental parameters in the nSCS, the time series of various parameters was determined. The parameters include physical parameters, such as SST, *T_{av}*, *T_{200m}*, SSS, *S_{av}*, *S₂₀₀*, and chemical parameters, such as SSDO, *DO_{av}*, *DO₂₀₀*, SSP, *P_{av}*, *P₂₀₀*, SSSi, *Si_{av}*, *Si₂₀₀*, SSDIN, *DIN_{av}*, *DIN₂₀₀*, SSNO₂-N, *NO₂-N_{av}*, *NO₂-N₂₀₀*, SSNO₃-N, *NO₃-N_{av}*, *NO₃-N₂₀₀*, SSNH₄-N, *NH₄-N_{av}*, *NH₄-N₂₀₀*, and the ratios of the chemical parameters SSN:SSP, *N_{av}:P_{av}*, *N₂₀₀:P₂₀₀*, SSSi:SSN, *Si_{av}:N_{av}*, *Si₂₀₀:N₂₀₀*, where av = average for the whole water column and 200=200 m depth. Statistical test and linear regression

analyses were conducted on time series (Chen and Ma, 1991) and climate trend coefficients (*R_{xt}*) were estimated. *R_{xt}* was used to assess whether there was a significant linear climate-trend in a time series (Shi et al., 1995). This coefficient was defined as the correlation coefficient between the time series of an environmental parameter, {*X_t*}, and the nature number {*i*}, *i*=1, 2, 3 ..., *n*. In this study, *n* is the total span of the years covered by the data. The coefficient was computed from the following equation:

$$R_{xt} = \frac{\sum_{i=1}^n (x_i - \bar{x})(i - \bar{i})}{\sqrt{\sum_{i=1}^n (x_i - \bar{x})^2 \sum_{i=1}^n (i - \bar{i})^2}} \tag{2}$$

Where *t*=(*n*+1)/2. Its significance level is determined from the Student *t*-test. A positive/ negative value of *R_{xt}* indicates that the time series, {*X_t*}, has a linear positive/negative trend. In order to compare the environmental change rates between coastal/shelf waters with the slope/open sea, data for water depths <200 and >200 m, respectively, were analyzed.

Biological oceanography data, such as chlorophyll-*a*, phytoplankton abundance, primary production, zooplankton biomass, benthos biomass, cephalopod catch, etc. were obtained during the period by marine ecosystem surveys conducted by the South China Sea Fisheries Research Institute (SCSFRI), the South China Sea Institute of Oceanography, Chinese Academy of Sciences and the Second Institute of Oceanography, SOA. For the observational methods, Chl-*a* was determined by acetone extraction fluorescence method (Holm-Hansen et al., 1965) and often calibrated by spectrophotometry. Primary productivity was measured using the ¹⁴C tracer method established by Steemann-Nielsen

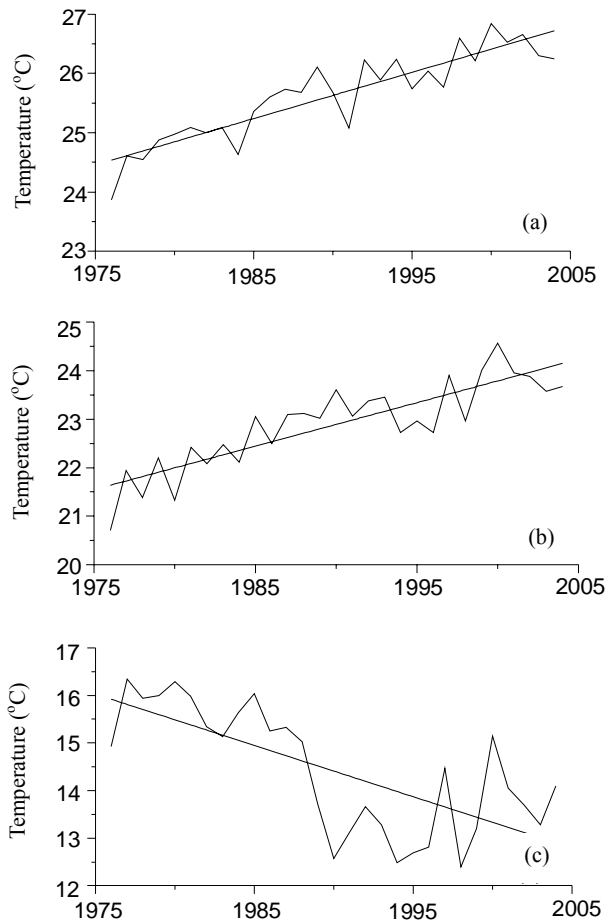


Fig. 2. Variation trends of seawater temperature in the nSCS during 1976–2004. (a), (b) and (c) show the annual mean of the sea surface temperature (SST), the water column average temperature in the upper 200 m (T_{av}), and the temperature at the depth of 200 m (T_{200}). The lines are the linear regressions.

(1952) and modified for scintillation counting by Wolfe and Schelske (1967). Phytoplankton samples were collected by vertical haul using a Judy net with a mesh size of $76\ \mu\text{m}$. The samples were preserved with Lugol's solution, and the species identification and cell counts were made using a microscope to get phytoplankton abundance (PA, Sournia, 1978). Zooplankton samples were collected by vertical haul using a plankton net with a mesh size of $505\ \mu\text{m}$, and the samples were preserved with neutral formaldehyde solution (5%). The species identification and individual counts were made using a stereo microscope, and the wet weight biomass (ZB) was measured by an electronic balance after removing the body surface water in the lab, according to SOAC (1975) and NBTS (1991). Benthic macrofauna samples were collected by using a grab with a sampling area of $0.1\ \text{m}^2$, and the animals were sorted after removing the mud by elutriation. The species identification and individual counts were made using a stereo microscope, and the wet weight biomass

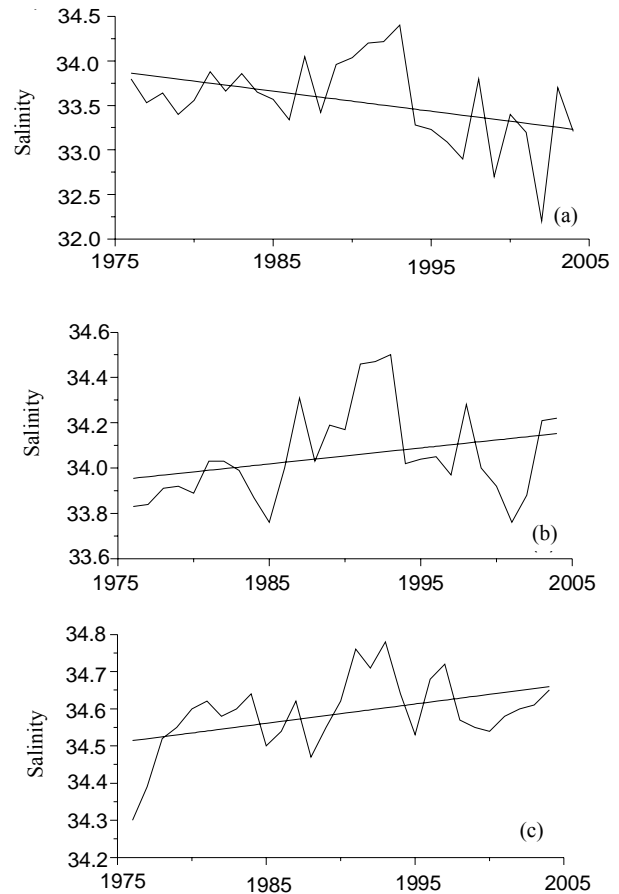


Fig. 3. Variation trends in seawater salinity in the nSCS during 1976–2004. (a), (b) and (c) show the annual means of sea surface salinity (SSS), water column average salinity in the upper 200 m (S_{av}) and salinity at 200 m (S_{200}), respectively. The lines are linear regressions.

(BB) was measured gravimetrically after removing the body surface water in the lab (SOAC, 1975; NBTS, 1991). The nekton samples were collected by using a cystoid net with a mesh size of 20 mm, towed by a pair of boats at a speed of 3–4 kn for 1 h at each station (SOAC, 1975; NBTS, 1991).

3 Results

3.1 Environmental changes during the 29 year time series

3.1.1 Seawater temperature and salinity

The annual means of SST and T_{av} (the average temperature in the water column) exhibited highly significant increasing trends ($p \leq 0.01$), their annual rates were 0.078 and $0.090\ ^\circ\text{C}\ \text{y}^{-1}$, respectively. However, the temperature at 200 m (T_{200}) showed a significant decreasing trend with an annual rate of $-0.108\ ^\circ\text{C}\ \text{y}^{-1}$ during the 29 years of observation between

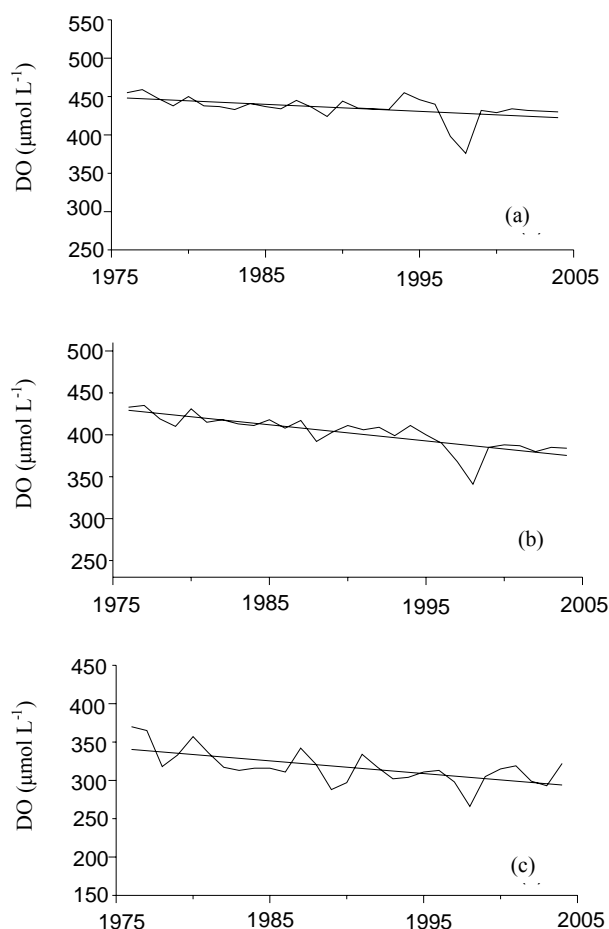


Fig. 4. Variation trends in dissolved oxygen (DO) concentration in the nSCS during 1976–2004. (a), (b) and (c) show the annual means of sea surface DO (SSDO), water column average in the upper 200 m DO (DO_{av}) and DO at the 200 m (DO_{200}), respectively. The lines are linear regressions.

1976 and 2004 (Fig. 2, Table 1). During this period, the SST and T_{av} increased by 2.26 and 2.61°C, respectively, while T_{200} decreased by 3.10°C (Fig. 2, Table 1). The R_{xt} of the time series of SST, T_{av} and T_{200} were 0.89, 0.85 and -0.71 , respectively, and the variation trends for this time series were significant ($p \leq 0.01$).

During the observation period the annual rate of SSS exhibited a decreasing trend (-0.022 y^{-1}), while the annual rates of S_{av} and S_{200} exhibited increasing trend (0.007 y^{-1} and 0.005 y^{-1} , respectively, Fig. 3, Table 1). The SSS decreased by 0.653, while the S_{av} and S_{200} increased by 0.206 and 0.151. The R_{xt} values of time series of SSS, S_{av} and S_{200} were -0.41 ($P \leq 0.05$), 0.38 ($P \leq 0.05$) and 0.44 ($p \leq 0.02$), respectively (Table 1).

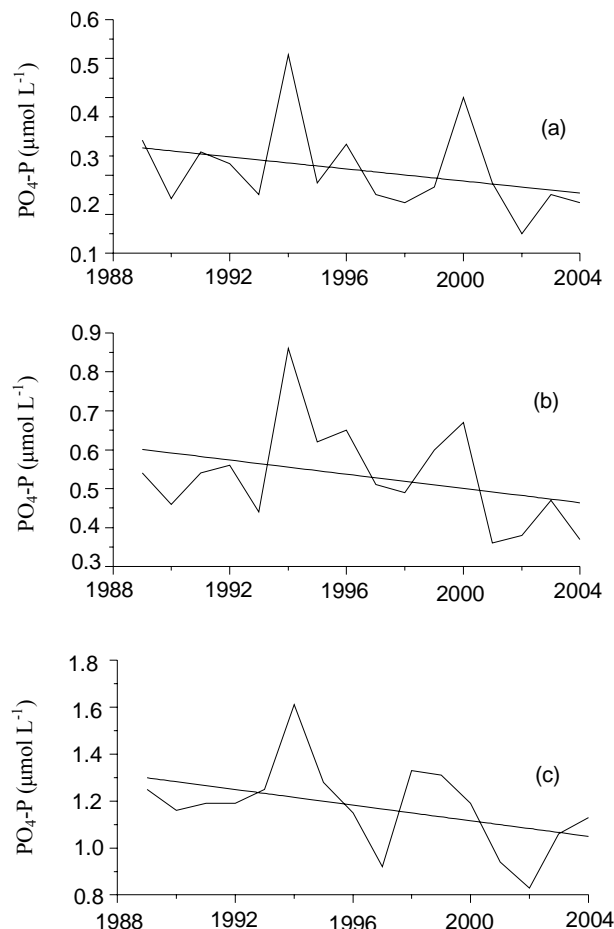


Fig. 5. Variation trends in P concentrations in the nSCS during 1989–2004. (a), (b) and (c) show the annual means of sea surface P (SSP), water column average P in the upper 200 m (P_{av}) and P in the 200 m layer (P_{200}), respectively. The lines are linear regressions.

3.1.2 Concentration of dissolved oxygen (DO)

A significant ($p \leq 0.01$) decrease in DO concentrations was observed (Fig. 4; Table 1), and the annual rate was between -0.91 and $-1.93 \mu\text{mol L}^{-1} \text{ y}^{-1}$. The regional mean of DO of the sea surface, the water column average in the upper 200 m and at 200 m significantly decreased by 26.5, 55.8 and $47.9 \mu\text{mol L}^{-1}$. The R_{xt} values of time series for DO ranged -0.48 to -0.81 ($p \leq 0.01$, Table 1).

3.1.3 Concentrations of P, Si and N

Both P and Si concentrations exhibited decreasing trends (Figs. 5 and 6, Table 1). Their annual rates were -0.008 to -0.017 (for P) and -0.071 to -0.387 (for Si) $\mu\text{mol L}^{-1} \text{ y}^{-1}$, respectively.

The concentrations of $\text{NO}_3\text{-N}$, $\text{NO}_2\text{-N}$, $\text{NH}_4\text{-N}$ and dissolved inorganic nitrogen ($\text{DIN} = \text{NO}_3 + \text{NO}_2 + \text{NH}_4$) in the nSCS exhibited obvious increasing trends during 1989–2004

Table 1. The annual rate, climate trend coefficients (R_{xt}) and the amplitude of fluctuations of the environmental parameters in the nSCS during 1976–2004 (units: °C y^{-1} for annual change in temperature, and $\mu\text{mol L}^{-1} y^{-1}$ for DO and nutrients).

	Annual rate			R_{xt}^a			Amplitude of fluctuation ^b			Mean \pm SD ^c		
	Mean <200 m	Mean >200 m	Mean whole area	Mean <200 m	Mean >200 m	Mean whole area	Mean <200 m	Mean >200 m	Mean whole area	Mean <200 m	Mean >200 m	Mean whole area
SST	0.078	0.063	0.078	0.84 ^A	0.80 ^A	0.89 ^A	2.67	3.28	2.97	24.75 \pm 1.82	26.06 \pm 0.67	25.63 \pm 0.74
T_{av}	0.089	0.095	0.090	0.80 ^A	0.85 ^A	0.85 ^A	4.07	4.41	3.86	23.49 \pm 0.97	22.42 \pm 0.95	22.29 \pm 0.89
T_{200}^d	0.045	-0.107		0.45 ^B	0.71 ^A		2.90	3.94		20.36 \pm 0.86	15.41 \pm 1.30	
SSS	-0.007	-0.015	-0.022	-0.19	-0.35	-0.41 ^C	3.61	2.30	2.20	33.10 \pm 0.78	34.01 \pm 0.40	33.55 \pm 0.48
S_{av}	0.008	0.005	0.007	0.32	0.34	0.38 ^C	0.97	0.71	0.74	33.89 \pm 0.30	34.22 \pm 0.19	34.05 \pm 0.21
S_{200}^d	0.005	0.005		0.30	0.44 ^B		0.42	0.48		34.34 \pm 0.20	34.59 \pm 0.10	
SSDO	-0.964	-0.927	-0.913	-0.46 ^B	-0.52 ^A	-0.48 ^A	80	85	83	426.5 \pm 76.0	425.4 \pm 15.0	434.6 \pm 15.9
DO_{av}	-1.857	-2.037	-1.926	-0.67 ^A	-0.82 ^A	-0.81 ^A	102	92	94	422.6 \pm 33.0	384.9 \pm 21.0	402.3 \pm 20.3
DO_{200}^d	-2.230	-1.653		-0.70 ^A	-0.63 ^A		99	104		376.3 \pm 28.0	317.2 \pm 22.4	
SSP	-0.011	-0.004	-0.008	-0.41	-0.30	-0.31	0.54	0.76	0.45	0.31 \pm 0.14	0.32 \pm 0.11	0.32 \pm 0.12
P_{av}	-0.008	-0.007	-0.009	-0.29	-0.35	-0.37	0.32	0.63	0.50	0.39 \pm 0.12	0.69 \pm 0.17	0.52 \pm 0.13
P_{200}^d	-0.006	-0.017		-0.22	-0.50 ^D		0.48	0.78		0.57 \pm 0.17	1.19 \pm 0.21	
SSSi	-0.642	-0.284	-0.071	-0.19	-0.30	-0.22	12.19	12.90	11.50	9.93 \pm 3.46	10.33 \pm 4.81	10.20 \pm 3.71
Si_{av}	-0.284	-0.240	-0.161	-0.24	-0.19	-0.16	17.84	21.46	17.86	11.51 \pm 6.05	15.15 \pm 7.40	12.64 \pm 7.17
Si_{200}^d	-0.038	-0.387		-0.06	-0.19		17.57	32.50		14.98 \pm 5.16	26.80 \pm 9.53	
SSDIN	0.613	0.391	0.504	0.70 ^A	0.63 ^A	0.74 ^A	17.27	9.69	11.85	4.44 \pm 4.24	3.73 \pm 3.00	4.09 \pm 3.29
DIN_{av}	0.416	0.343	0.381	0.75 ^A	0.61 ^A	0.73 ^A	9.68	10.53	10.53	4.71 \pm 2.63	6.33 \pm 2.65	6.71 \pm 2.60
DIN_{200}^d	0.721	0.813		0.73 ^A	0.64 ^A		15.71	20.13		7.61 \pm 4.68	16.08 \pm 6.03	
SSNO ₃	0.457	0.342	0.456	0.70 ^A	0.54 ^D	0.66 ^A	12.05	9.56	9.74	2.62 \pm 3.09	2.45 \pm 2.99	2.55 \pm 2.85
NO_{3av}	0.345	0.324	0.335	0.71 ^A	0.58 ^B	0.67 ^A	8.46	9.85	8.13	2.98 \pm 2.31	5.09 \pm 2.68	4.01 \pm 2.37
NO_{3200}^d	0.582	0.711		0.68 ^A	0.55 ^C		13.57	20.55		5.57 \pm 4.05	14.53 \pm 6.17	
SSDIN:SSP	3.61	1.89	2.75	0.55 ^C	0.54 ^D	0.63 ^A	126.55	52.65	84.02	20.5 \pm 30.7	14.9 \pm 16.8	17.8 \pm 21.2
$DIN_{av}:P_{av}$	1.66	0.75	1.21	0.70 ^A	0.57 ^B	0.69 ^A	40.26	23.24	28.65	14.0 \pm 11.3	10.7 \pm 6.3	12.3 \pm 8.4
$DIN_{200}:P_{200}^d$	1.51	1.04		0.74 ^A	0.64 ^A		31.26	27.3		13.8 \pm 8.6	14.0 \pm 7.8	
SSSi:SSDIN	-0.52	-0.46	-0.44	-0.75 ^A	-0.39	-0.67 ^A	12.32	22.68	9.69	4.1 \pm 3.3	5.4 \pm 5.7	4.3 \pm 3.2
$Si_{av}:DIN_{av}$	-0.46	-0.15	-0.22	-0.71 ^A	-0.46	-0.70 ^A	13.05	5.62	5.41	3.4 \pm 3.1	2.9 \pm 1.6	2.8 \pm 1.5
$Si_{200}:DIN_{200}^d$	-0.32	-0.10		-0.75 ^A	-0.65 ^A		7.29	2.8		2.87 \pm 2.02	1.85 \pm 0.8	

^a Significance: ^A $p \leq 0.01$, ^B $p \leq 0.02$, ^C $p \leq 0.03$, and ^D $p \leq 0.05$

^b The amplitude is the difference between maximal annual mean value and minimal one during the observation period. It indicates quantitative characteristics of the fluctuation of a parameter during the observation period. Standard deviation is symbolized as $X_{\delta n-1}$.

^c Mean \pm SD is the multi-year's mean value and standard deviation during the observation period.

^d When the water depth <200 m, data for the bottom layer are used; when the water depth >200 m, data for 200 m were used.

(Fig. 7, Table 1). The annual rates of the increases in DIN and NO₃-N were 0.38 to 0.81 and 0.34 to 0.71 $\mu\text{mol L}^{-1} y^{-1}$, respectively. The regional means of DIN values of the sea surface, water column average and 200 m layer increased by 8.06, 6.10 and 13.01 $\mu\text{mol L}^{-1}$, respectively. DIN concentrations in the 200 m layer were significantly higher than those of the average for the water column and at the sea surface. Its annual rate was 1.61 times greater than that of the sea surface. And its multi-year average value during observation period was 3.9 times higher than that of the sea surface layer. The R_{xt} value for the time series of NO₃-N and DIN were between 0.55–0.74. The variation trends of the NO₃-N and DIN time series were highly significant ($p \leq 0.01$) (Table 1), except for the time series of NO₃-N₂₀₀, which was significant ($p \leq 0.03$). It was also found that the annual rate of DIN was higher in the shallow (<200 m) than in the deep water (>200 m) areas (Table 1). The spatial-temporal distributions of the DIN_{av} (mean concentrations in the water column) indicated that since 1998 the high DIN_{av} concen-

trations have exhibited a pronounced large area increasing, i.e. after 1998 in the most study area, DIN_{av} concentration >5.71 $\mu\text{mol L}^{-1}$ (Fig. 8, the shadow shows the area), and exceeded the low limit of suitable N concentration for diatom growth (5.71 $\mu\text{mol L}^{-1}$, Chu, 1949). So it is favorable for phytoplankton growth in the most study area.

3.1.4 Nutrient ratios

During 1989–2004, the nutrient ratios (DIN:P and Si:DIN) changed significantly. For example, the DIN:P ratios increased by 1.04–2.75 y^{-1} , and it reached 28.1 (SSDIN:SSP) and 18.0 ($DIN_{av}:P_{av}$) in 2004 (Fig. 9, Table 1); the Si: N ratios decreased at an annual rate of $-0.10 \sim -0.44 y^{-1}$, and it dropped to 1.0 (SSSi:SSDIN) and 1.4 ($Si_{av}:DIN_{av}$) in 2004 (Fig. 10, Table 1). The variation trends of all time series of N: P and Si: N were significant ($p \leq 0.01$). The annual rates of increase in DIN:P and decrease in Si:DIN were higher in the shallow coastal waters (<200 m) than in deep continental shelf waters (>200 m) (Table 1).

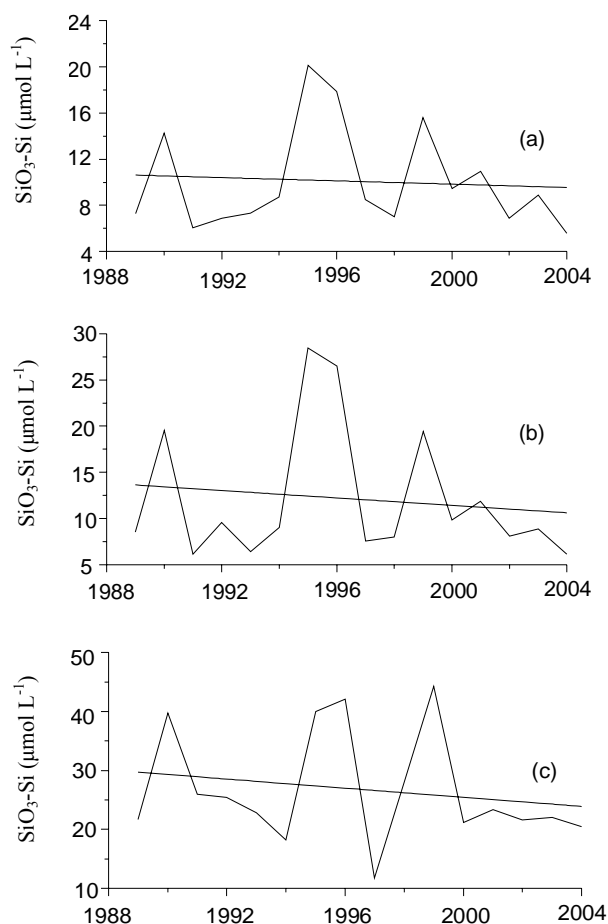


Fig. 6. Variation trends in Si concentrations in the nSCS during 1989–2004. (a), (b) and (c) show the annual means of sea surface Si (SSSi), water column average Si in the upper 200 m (Si_{av}) and Si in the 200 m layer (Si_{200}), respectively. The lines are linear regressions.

4 Discussion

4.1 Increasing trends and the response of the ecosystem

The positive increasing trends in SST and T_{av} in the nSCS during 1976–2004 are consistent with the increasing trends in the mean air temperature (AT) observed throughout the Northern Hemisphere (Houghton et al., 1996; Fu et al., 2006), South China (Chen et al., 1998; Chen et al., 1999; Zhai and Ren, 1997) and the annual means of AT and SST observed along the coast of the SCS (He et al., 2003; Martin and Arun, 2003). The increasing trends were also in phase with the changes in SST observed along the coast of the Yellow Sea and Bohai Sea (Lin et al., 2005, 2001). However, these annual rates of water temperature change were higher in the nSCS than in the Yellow Sea and the Bohai Sea (Lin et al., 2005, 2001).

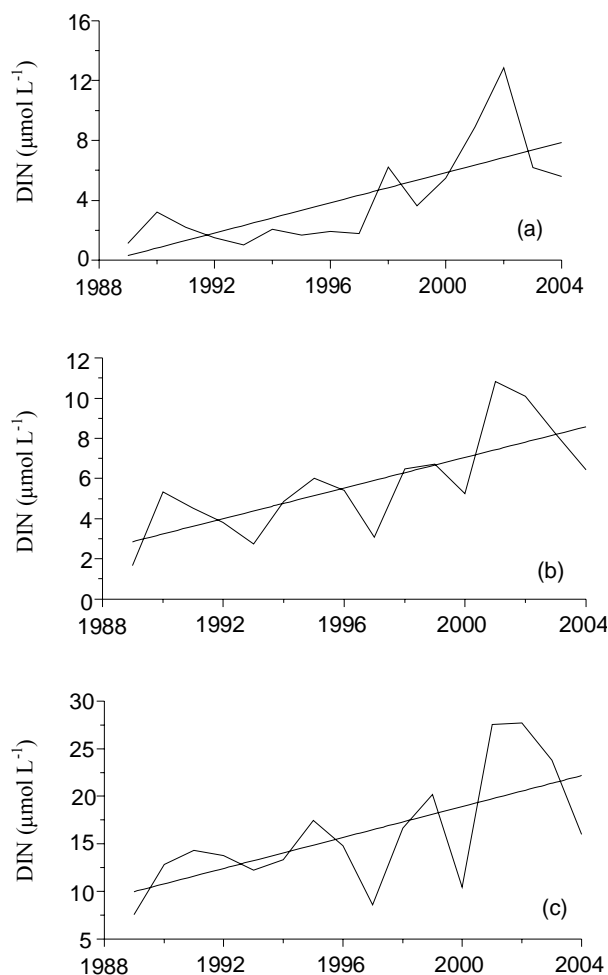


Fig. 7. Variation trends in dissolved inorganic nitrogen (DIN) concentrations in the nSCS during 1989–2004. (a), (b) and (c) show the annual means of sea surface DIN (SSDIN), water column average DIN in the upper 200 m (DIN_{av}) and DIN at the 200 m layer (DIN_{200}), respectively. The lines are linear regressions.

The increase in DIN in the nSCS during 1989–2004 was consistent with the increase in DIN along the coast of the nSCS, such as Qinzhou Bay (Wei et al., 2002, 2003) and Daya Bay (Qiu et al., 2005). It also shows the same trend with the rise of DIN observed throughout the global marginal seas (Seitzinger et al., 2002). Along with the rapid economic development in China, DIN concentration in the Pearl River estuary and shelf of the China Sea has been dramatically increased, due to the increasing urbanization near the coastal areas, which resulted in more municipal sewage, agricultural fertilizer, mariculture waste etc. inputs (SOAC, 2001). Through analysis based on DIN and PN models, combining with spatially explicit global databases, Seitzinger et al. (2002) showed that DIN input rates increased from approximately 21 Tg N y^{-1} in 1990 to 47 Tg N y^{-1} by 2050. The largest increases are predicted for Southern and Eastern

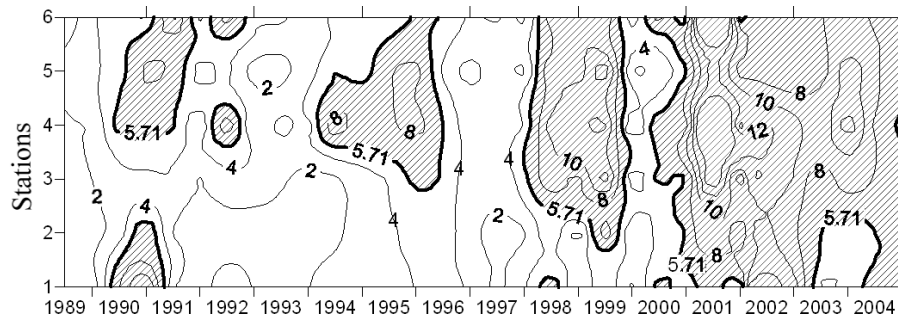


Fig. 8. The spatial-temporal distributions of the DIN_{av} (mean concentrations in the water column). The shadow shows the area and period of time which exceeded the low limit of suitable N concentration for diatom growth ($5.71 \mu\text{mol L}^{-1}$, Chu, 1949).

Table 2. Changes in annually average concentration ($\mu\text{mol dm}^{-3}$) of $\text{NO}_3\text{-N}$ and dissolved inorganic nitrogen (DIN), from discharges of the Pearl River and other rivers.

Observation year	Location of the observation	Parameter, value	Data source
1980–1985	southern Hong Kong waters (outside Pearl River estuary)	$\text{NO}_3\text{-N}$, 1–3	Han et al. (1990)
1986	Pearl River estuary	DIN, 19.3	He et al. (2004)
1987	Pearl River estuary	$\text{NO}_3\text{-N}$, 28.0 DIN, 31.7	Wang and Peng (1996)
1990	Pearl River estuary	DIN, 34.6	He et al. (2004)
1995	Pearl River estuary	DIN, 36.4	He et al. (2004)
1996	Pearl River estuary	$\text{NO}_3\text{-N}$, 39.8	Cai (2002b)
1999 (summer)	Pearl River estuary	$\text{NO}_3\text{-N}$, 48.9 DIN, 51.9	Lin et al. (2004)
1999 (summer)	Pearl River estuary	DIN, 53.2	Guan et al. (2003)
2002	Pearl River estuary	DIN, 76.4	He et al. (2004)

Asia, associated with predicted large increases in population, will increase fertilizer use, and increased industrialization. DIN from the Pearl River discharge increased by 3 times in 2002 as compared to 1986 (He et al., 2004, Table 2), and $\text{NO}_3\text{-N}$ input from the Pearl in River Estuary was 1.7 times in 1999 of that in 1987 (Guan et al., 2003; Wang and Peng, 1996, Table 2). SSDIN should be influenced by the increase in DIN from the river discharges. In addition, significant inputs of DIN into the nSCS have also occurred through atmospheric dry and wet deposition (Zhang et al., 1999) and the upwelling of the deep waters (Zhao et al., 2005). The mitigation of N limitation in the upper layer since 1998 was clearly related to these DIN inputs (Fig. 8). The increase in the annual rate of DIN was higher at the 200 m layer (DIN_{200}) than the water column average (DIN_{av}) and at the sea surface layer (SSDIN). This may be due to possible strengthening of the deep water upwelling. However, during 1989–1997, the DIN_{av} was low, the multi-annual mean was $4.2 \mu\text{mol L}^{-1}$ and the lowest value of DIN_{av} was only $1.7 \mu\text{mol L}^{-1}$ in 1989. Since 1998, the multi-annual mean of the water column average DIN concentration has reached $7.7 \mu\text{mol L}^{-1}$ and the maximal value of the annual mean even reached up to $10.8 \mu\text{mol L}^{-1}$ in 2001 (Fig. 7b). And therefore, the previous status of N limitation in the nSCS was significantly al-

leviated. Moreover, the increases in the annual rates of DIN and $\text{NO}_3\text{-N}$ were much higher in the nSCS than in the Yellow Sea (Lin et al., 2005).

The increase in the N:P ratio was due to an increase in DIN and a decrease in P concentration. Before 1997, the N:P ratios (SSN:SSP, $DIN_{av}:P_{av}$ and $DIN_{200}:P_{200}$) were lower than 10. Since 1998, these ratios have rapidly increased to 28, 18 and 16 in 2004, respectively (Fig. 9). In 2004, the average values of $DIN_{av}:P_{av}$ and $DIN_{200}:P_{200}$ were close to the Redfield ratio (16:1), and therefore favorable to phytoplankton growth (Richardson, 1997; Hu et al., 1989; Jiang and Yao, 1999). The high value of SSN:SSP (28) in 2004 was probably due to dry and wet deposition. The peak value of SSN:SSP appeared in 2002 (up to 86, Fig. 8a), due mainly to high rainfall as high Pearl River discharge, and this high ratio corresponded to the lowest surface salinity (Fig. 3a). Moreover, the fact that the annual rates of DIN and DIN:P were higher in the shallow water area (<200 m) than in the deep water area (>200 m) (Table 1) suggested the influence of the anthropogenic activities on the ecosystems of the shallow or coastal waters. These results agree well with Xu et al. (2008) who also showed that N input from the Pearl River has caused the nSCS to become P-limited. Phosphate decline may be ascribed to the implementation of phosphate

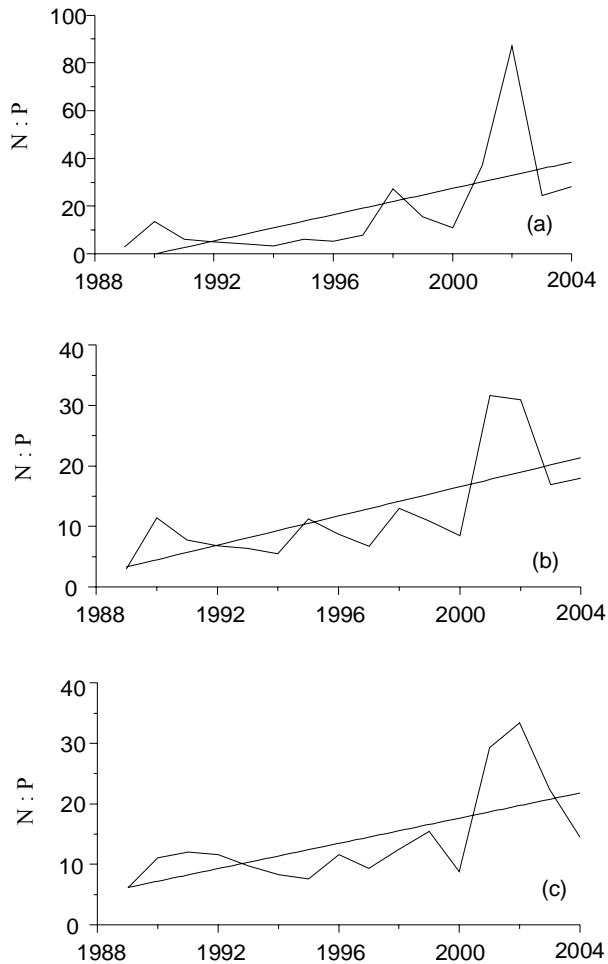


Fig. 9. Variation trends in N:P ratios in the nSCS during 1989–2004. (a), (b) and (c) show the annual means of sea surface DIN:P (SSDIN:SPP), water column average DIN:P in the upper 200 m ($DIN_{av}:P_{av}$) and DIN:P at the 200 m layer ($DIN_{200}:P_{200}$), respectively. The lines are linear regressions.

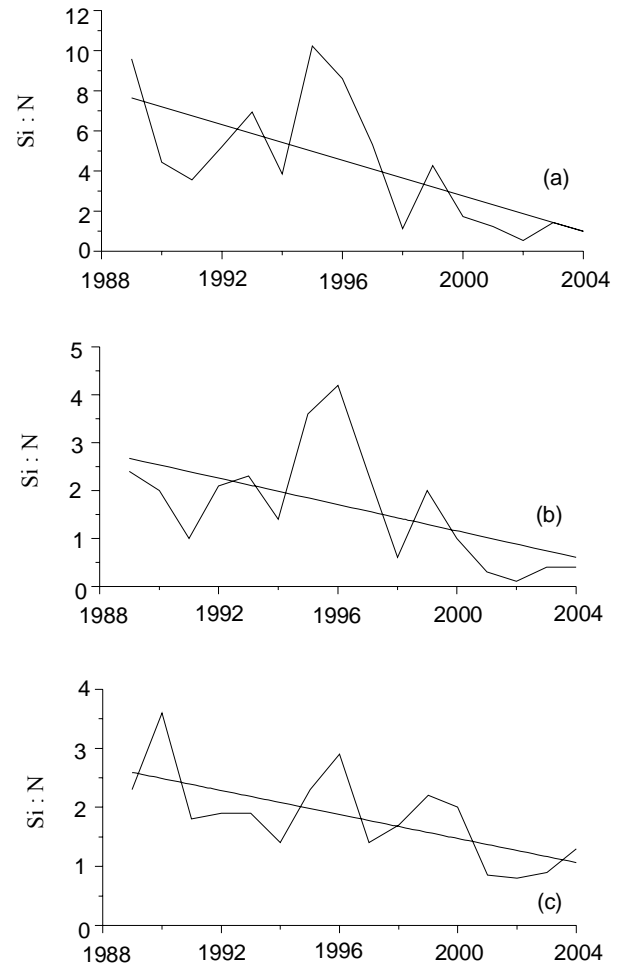


Fig. 10. Variation trends in Si:N ratios in the nSCS during 1989–2004. (a), (b) and (c) show the annual means of sea surface Si:N (SSSi: SSDIN), water column average Si:N in the upper 200 m ($Si_{av}:DIN_{av}$) and Si:N at the 200 m layer ($Si_{200}:DIN_{200}$), respectively. The lines are linear regressions.

detergent ban in the late 1990s, and phytoplankton great consumption.

4.2 Decreasing trends and the response of the ecosystem

The decreasing trends in SSS may be related to the increase in the freshwater discharge into the nSCS since 1990 and less vertical mixing, because of the presence of a permanent thermocline (Yuan and Deng, 1997a, b; Shi, 2001). It has previously been observed that the occurrence of low SSS often corresponded to abnormally high discharge of the Pearl River (Xie and Zhang, 2003; Xu et al., 2008). During the period of 1990 to 2000, the Pearl River runoff increased by 22.5% in comparison with the mean discharge for 1960–1999 (Lei et al., 2003). In the 1990s (i.e. 1994, 1995, 1996, 1997, 1999) and 2002, SSS in the nSCS was especially low (Fig. 3a), which was induced by the Pearl River floods (Lei

et al., 2003). Particularly in 2002, the strongest typhoon rain storm resulted in the historically greatest flood of the Pearl River, and therefore, which induced in the lowest SSS (32.2) in the nSCS during that study period (Lei et al., 2003; Xie and Zhang, 2003). In 1999, the SSS was below normal due to the influence of frequent typhoon rain storms (total of 28 typhoons, including 7 landed ones). Furthermore, in the summer of 1999, the Dongsha upwelling (Station 3–6) was weak and the depth of the 20°C isopleth was deeper than 105 m. However in 1998, although large floods of the Pearl River occurred, the SSS did not decrease, due to the strong upwelling as indicated by shallow (84 m) depth of the 20°C isopleths. Chai et al. (2001) pointed out that the strength of upwelling can be indicated by the depth of 20°C temperature isopleth in the SCS. In this study, the range depth of 20°C isopleths is 84–150 m. However, the high SSS in 1993 was probably due to the reduced Pearl River discharge (Lei et al., 2003).

Table 3. Comparison of key nutrient concentration ($\mu\text{mol L}^{-1}$) and the ratio between the two phases (before 1997, the first phase and after 1997, the second phase).

Nutrient and ratio	mean in the first phase	mean in the second phase
SSDIN	1.83	5.75
DIN_{av}	4.22	7.78
DIN_{200}	12.77	20.33
SSP	0.33	0.25
P_{av}	0.56	0.48
P_{200}	0.49	0.48
SSSi	12.38	9.19
Si_{av}	14.82	12.26
Si_{200}	27.54	25.85
SSDIN:SSP	5.5	23.0
$\text{DIN}_{av}:\text{P}_{av}$	7.5	16.2
$\text{DIN}_{200}:\text{P}_{200}$	26.1	42.4
SSSi:SSDIN	6.3	1.6
$\text{Si}_{av}:\text{DIN}_{av}$	2.5	1.6
$\text{Si}_{200}:\text{DIN}_{200}$	2.2	1.3
SSSi:SSP	37.5	36.8
$\text{Si}_{av}:\text{P}_{av}$	26.5	25.5
$\text{Si}_{200}:\text{P}_{200}$	56.2	53.8

The decreasing trend in DO concentration (Fig. 4) can probably be attributed to three reasons. First, a small decrease in DO solubility in the seawater can be induced by the increase in the seawater temperature (Fig. 2). Second, an increase in DO consumption was an important reason. The DO consumption resulted from the decomposition of organic matter originating mainly from the Pearl River discharge and the decay of phytoplankton blooms with increasing frequencies in the coastal water was increasing (Peng, 1994; Tang et al., 2006). Before 1998, HABs (harmful algae blooms) occurred once or twice a year, but during 1998 to 2003, blooms increased to 10 to 20 a year in the nSCS (Tang et al., 2006), which was in phase with the dramatic decreasing trend in DO (Fig. 4). Third, the mixing between the surface water and the deep layers was reduced by the stronger thermocline, due to the rapid rise in SST since 1995 (Fig. 2a), which resulted in less transfer of oxygen from the atmosphere to the deeper waters. In addition in 1998, the lowest value of DO was probably attributed to the strongest upwelling, which occurred that year. According to the compute of this study, in 1998 the depth of the 20°C isopleths was 84 m, it show the strongest upwelling (see above paragraph and references in it).

The decrease in P concentration is probably due to uptake by phytoplankton and less P supply from deep water, due to the presence of the permanent thermocline (Yuan and Deng, 1997a, b; Shi et al., 2001). Furthermore, since 1998, the reduction in N limitation has increased the phytoplank-

ton biomass and production (Table 3) and decreased the P concentration. Hong et al. (1983) reported that diatoms can take up 30 times more P than they need and store it for use when P is deficient. In the most years except for 1994 and 2000, the concentrations of P in the upper layer (<75 m) were even lower than the P concentration required for diatom growth ($<0.1 \mu\text{mol L}^{-1}$, Zou et al., 1983). The decreasing trends in Si concentration have probably been influenced by the decrease in Si concentration in the Pearl River runoff since the 1970s (Lei et al., 2003), but still adequate for diatom growth. At 200 m, the interannual fluctuations of both P and Si concentrations were high (Figs. 5 and 6) and probably attributed to the interannual changes in upwelling of deep water. In 1996 and 1999 the depth of the 20° isopleth was shallow (average value from multi-stations was 93 m), high concentrations of P and Si occurred ($\text{P}_{av}=0.61$ and $0.62 \mu\text{mol L}^{-1}$, $\text{P}_{200}=1.154$ and $1.31 \mu\text{mol L}^{-1}$, $\text{Si}_{av}=25.18$ and $19.55 \mu\text{mol L}^{-1}$, and $\text{Si}_{200}=42.12$ and $44.28 \mu\text{mol L}^{-1}$, respectively, in the two years). In contrast, during 2004 when the depth of the 20° isopleth was deep ((the average value from multi-stations was 105 m), concentrations of P ($\text{P}_{av}=0.37 \mu\text{mol L}^{-1}$, and $\text{P}_{200}=1.13 \mu\text{mol L}^{-1}$) and Si ($\text{Si}_{av}=8.9 \mu\text{mol L}^{-1}$, and $\text{Si}_{200}=20.5 \mu\text{mol L}^{-1}$) were obviously low.

4.3 Nutrient limitation

Since 1998, DIN concentrations have exhibited a pronounced increasing trend, and therefore the key nutrient concentration and ratio can be divided into two phases, i.e. before and after 1997 (Table 3). In the first phase, the average DIN concentration was very low, SSDIN and DIN_{av} were 1.83 and $4.22 \mu\text{mol L}^{-1}$, respectively, which were lower than the low limit of suitable N concentration for diatom growth ($5.71 \mu\text{mol L}^{-1}$, Chu, 1949). The average P concentration was also low. SSP and P_{av} were 0.33 and $0.56 \mu\text{mol L}^{-1}$, respectively, which is closed to the low limit of suitable P concentration for diatom growth ($0.48 \mu\text{mol L}^{-1}$, Zhao et al., 2000). However, Si concentrations (SSSi and Si_{av} were 12.38 and $14.82 \mu\text{mol L}^{-1}$, respectively) outclassed the low limit of suitable for diatom growth ($4.40 \mu\text{mol L}^{-1}$, Harvey, 1957). Therefore, Si pool was sufficient. N:P ratios were less than half the Redfield ratio of 16 (SSN:SSP and $\text{N}_{av}:\text{P}_{av}$ were 5.5 and 7.5, respectively).

In contrast, for the second phase after 1997, the average DIN concentration has been clearly increasing (SSDIN and DIN_{av} were 5.75 and $7.78 \mu\text{mol L}^{-1}$ respectively, Table 3), and exceeded the low limit of suitable N concentration for diatom growth ($5.71 \mu\text{mol L}^{-1}$, Chu, 1949). However, the average P concentration has decreased (SSP and P_{av} , were 0.25 and $0.48 \mu\text{mol L}^{-1}$, respectively, Table 3). Consequently, there was a rapid increase in DIN:P ratios (SSDIN:SSP and $\text{DIN}_{av}:\text{P}_{av}$ were 23.0 and 16.2, respectively, Table 3), which were close to the Redfield ratio (16, Richardson, 1997; Hutchins et al., 1998). Therefore, it is favorable to

phytoplankton growth in the second phase. Si concentration also decreased ($SSSi$ and Si_{av} were 9.19 and 12.26) in this phase, resulting in rapid decrease in the ratio of $SSSi:SSDIN$ ($Si_{av}:DIN_{av}$), from 6.3 (2.5) to 1.6 (1.6) (Table 3). The Si concentration decrease may be due to the dam of the river courses which significantly reduced the silica delivery to the SCS.

Based on the studies on the kinetics of nutrient uptake, the thresholds of SiO_3-Si , DIN and PO_4-P for phytoplankton growth have been estimated to be 2.0, 1.0 and $0.1 \mu\text{mol L}^{-1}$, respectively (Justic et al., 1995). In the study area, the values of all the nutrient parameters were over these threshold concentrations, except for those in the first phase, when the mean of $SSDIN$ was close to the threshold of N (Table 3). According to chemical stoichiometry, in the first phase both $SSN:SSP$ and $DIN_{av}:P_{av}$ was lower than 10, and $SSSi:SSDIN$, $Si_{av}:DIN_{av}$ and $Si_{200}:DIN_{200}$ were all over 1, indicating potential N limitation. In the second phase $SSDIN:SSP$ and $DIN_{200}:P_{200}$ were higher than 22, and $DIN_{av}:P_{av}$ was equal to the Redfield ratio. All $Si:DIN$ ratios ranged from 1.3 to 1.6, and $Si:P$ ranged 25.5 to 53.8 (Table 3), which indicated that the potential of P limitation increased and N limitation decreased. Furthermore, Si was always sufficient during the observation period, even in the second phase when its concentration decreased (Table 3).

4.4 Response of ecological environment to ENSO events

When El Nino occurs, the warm pool of the western Pacific Ocean moves eastward, whereas it moves westward during La Nina (White et al., 1985; Takeuchi, 1987; Zhang and Huang, 1993). The nSCS is located to the west of the warm pool – however the response of the nSCS has not been well documented. In the present study, pronounced responses to ENSO were found. During the observation period, 9 El Nino events (1976, 1982–1983, 1986–1987, 1991, 1993, 1994, 1997, 2002 and 2004) and 4 La Nina events (1981, 1988, 1995 and 1998–1999) occurred (Wang and Gong, 1999; Qin, 2003; Mcphaden, 2004; Levimson, 2005). In general, whenever El Nino/La Nina occurred, SST and T_{av} was low/high in the nSCS (Fig. 2, Table 4).

In the area southwest of Dongsha Islands, near Station 4 ($19\text{--}20^\circ\text{N}$, $116\text{--}117^\circ\text{E}$, Fig. 1), pronounced responses to ENSO were observed. In general, whenever an El Nino event occurred, T_{av} and DO were lower, and that the S_{av} , nutrients (PO_4-P_{av} , SiO_3-Si_{av} and DIN), sea surface $Chl-a$ were higher (Table 5, Fig. 11). During a La Nina event, contrary, T_{av} and DO were higher, and that the S_{av} , nutrients (PO_4-P_{av} , SiO_3-Si_{av} and DIN), sea surface $Chl-a$ were lower (Table 5, Fig. 11). The fluctuations of environmental parameters evident corresponded to El Nino/La Nina events. Furthermore, in comparison of the average values of the environmental parameters in summer of El Nino years with those of La Nina years, T_{av} and DO_{av} were lower by 1.89°C and $20.2 \mu\text{mol L}^{-1}$, respectively; while S_{av} was higher by

0.31 psu ; PO_4-P_{av} , SiO_3-Si_{av} and DIN_{av} were higher by 0.15, 6.41 and $3.42 \mu\text{mol L}^{-1}$, respectively. Surface $Chl-a$ concentration was higher by 0.14 mg m^{-3} , i.e. higher by 1.8 times (Table 5), even higher by 0.83 times than the average value of normal years (1980, 1990, 2000, 2001 and 2003, for which mean $Chl-a=0.12\pm 0.05 \text{ mg m}^{-3}$).

According to Takano et al. (1998) and Liao et al. (2006), there is a cyclonic eddy in the sea area around Station 4 (near Dongshan Islands) in summer. Whenever a medium and strong El Nino occurs, the summer monsoon is weak (Zhang et al., 2003; Zhu et al., 2000), which induces strengthened cyclonic eddy, leading to strong upwelling, resulting in low T_{av} and DO_{av} , and high S_{av} , nutrients and $Chl-a$ induced by phytoplankton growth. During a La Nina event, the opposite occurs. This is due to ENSO/La Nina events affecting the strength of the summer monsoon related, i.e. both El Nino and La Nina make the anomaly of Walker circulation. During El Nino event, the heat convection of the western Pacific warm pool moves to the central Pacific. In contrast, during La Nina event, the heat convection of the western Pacific warm pool moves back to the western Pacific. That makes the anomaly of Walker circulation. Namely, when phase of El Nino (La Nina) fall under the influence of the anomalous Walker circulation, subsidence (ascending) air current occur over low latitude and middle latitude of the east Asia. If this subsidence (ascending) air current flow over low latitude occurred in summer, and joined to west-south monsoon, it will result in weaken (strengthen) of the summer monsoon (Wang et al., 2001; Zhang et al., 2003; Zhu et al., 2000). The changes of ecological environment in the sea area around Station 4 in summer respond to ENSO events, namely, respond to the abnormality of summer monsoon.

4.5 Response of the ecosystem and living resources

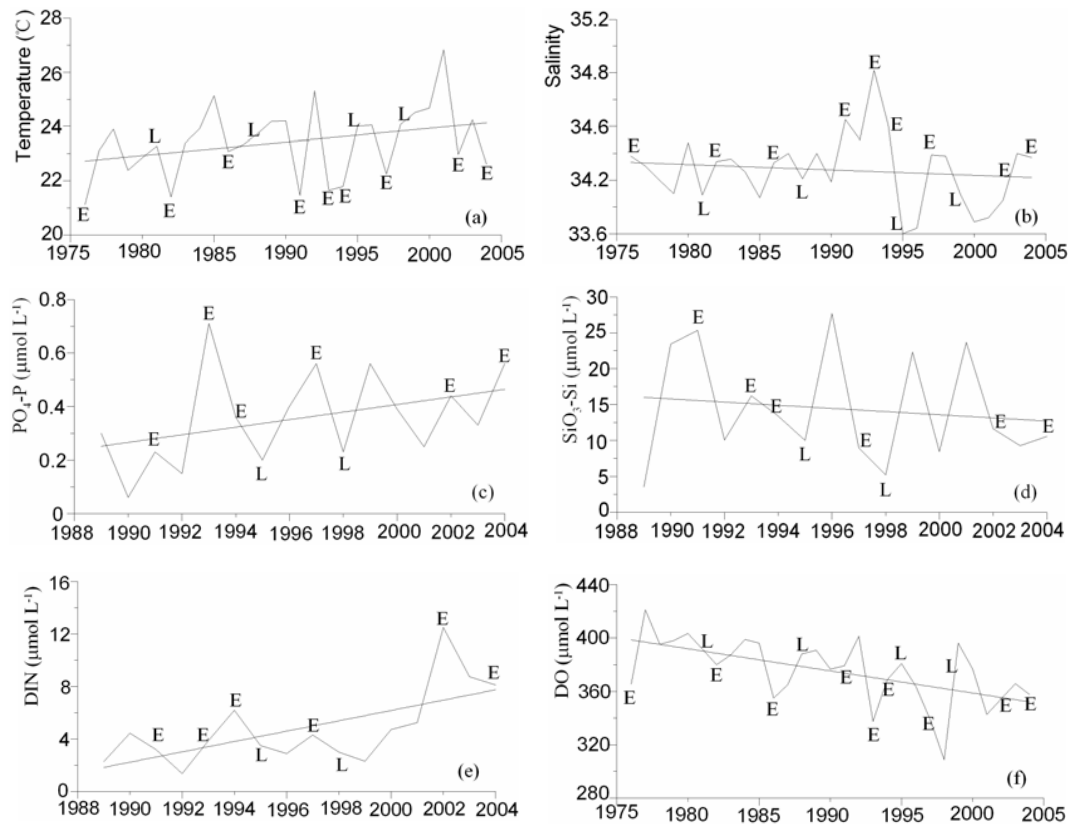
Although the frequency of observation on various biology and fisheries parameters was less than that for physical and chemical parameters which reflected the long term changes in the environmental processes, some responses of the ecosystems to these environmental changes in the nSCS was still evident.

Comparing the biology and fisheries data between the two phases discussed above, the average values of $Chl-a$, primary production, phytoplankton abundance, benthos biomass, cephalopod catch and demersal trawl catch increased by 6.4, 1.4, 2.4, 0.7, 7.2 and 2.8 times, respectively during the second phase, except for zooplankton abundance which decreased by about 50% (Table 6). The decrease in zooplankton abundance may be related to the increase in its predators, such as fish, cephalopods, etc. (Fig. 12). The R_{xt} values of the cephalopod and demersal trawl catches were 0.89 and 0.88, respectively, and highly significant ($p\leq 0.01$). Besides, the increase in both the cephalopod and demersal trawl catches partially could be attributed to the improvement of demersal trawl fishing techniques, and also to the increase

Table 4. Annual mean of SST and T_{av} along transect N in the nSCS in El Nino and La Nina years during 1976–2004 (unit: °C).

El Nino Year	1976	1982 (1983)	1986 (1987)	1991	1993	1994	1997	2002	2004	Mean±SD ^a
SST	23.87	25.82 (25.09)	25.60 (25.73)	25.08	25.89	26.24	25.77	26.66	26.25	25.64±0.75
T_{av}	21.13	21.40 (23.40)	23.09 (23.32)	21.46	21.65	21.79	22.25	22.98	22.62	22.27±0.83
La Nina Year		1981	1988	1995	1998	1999				
SST		25.09	25.68	25.74	26.61	26.21				25.87±0.58
T_{av}		23.28	23.72	24.00	24.07	24.51				23.92±0.46

^a Mean±SD is the multi-year mean value ± (SD), standard deviation for the El Nino/La Nina years.

**Fig. 11.** The interannual changes of the water column average values of the seawater temperature (a), (b) salinity, (c) $PO_4\text{-P}$, (d) $SiO_3\text{-Si}$, (e) DIN and (f) DO at the Station 4 of the nSCS in summer during the observation period.

in stock and production of low trophic levels, induced by the alleviation in N limitation in the nSCS.

In addition, after 1997, phosphorus depletion in surface waters during summer coincided with a shift in the dominant species in phytoplankton community from diatoms to dinoflagellates and cyanophytes (Ning et al., 2004). Peng et al. (2006) pointed out that dinoflagellates composed more than 60% of the total phytoplankton abundance in the HK Southeast Anti-Cyclonic Eddy and Hainan Island East Anti-Cyclonic Eddy of the nSCS, where P concentration was near detection limit in the summer of 2004.

In second phase our above observations were consistent with the results obtained by the multidisciplinary investigations for assessing the environmental health status and the fisheries environment quality in the nSCS during 1997–2002 (Jia et al., 2005). They reported that both primary production and phytoplankton abundance were relatively high, averaging $409.7 \text{ mg C m}^{-2} \text{ d}^{-1}$, and $837 \times 10^3 \text{ cell m}^{-3}$, respectively (according to criterion of grade of primary productivity and diet organism lever of China, Jia et al., 2003). The benthic biomass (averaging 11.3 g m^{-2}) was at normal levels (the same as above row, Jia et al., 2003). The zooplankton biomass (averaging 22.1 mg m^{-3}) was low (the

Table 5. The mean of the ecological parameter for the water column (0–200 m) and sea surface Chl-*a* concentration at Station 4 in summer of El Nino/La Nina years during the study period (units: °C for temperature, psu for salinity, $\mu\text{mol L}^{-1}$ for DO and nutrients and mg m^{-3} for Chl-*a*).

El Nino Year	1976	1982 (1983)	1986 (1987)	1991	1993	1994	1997	2002	2004	Mean±SD ^a
T_{av}	21.13	21.40 (23.40)	23.09 (23.32)	21.46	21.65	21.79	22.25	22.98	22.62	22.26±0.80
S_{av}	34.18	34.14 (34.16)	34.13 (34.2)	34.45	34.82	34.41	34.19	33.85	34.17	34.25±0.25
DO_{av}	365.6	380.1 (387.3)	355.1 (364.9)	379.2	337.3	368.5	339.0	354.2	357.8	362.6±16.0
$\text{PO}_4\text{-P}_{av}$				0.23	0.71	0.36	0.56	0.44	0.56	0.48±0.17
$\text{SiO}_3\text{-Si}_{av}$				25.36	16.22	13.38	8.96	11.65	10.56	14.36±5.94
DIN_{av}				3.20	3.85	6.19	4.29	12.51	8.12	6.36±3.50
Chl- <i>a</i> ^b		0.31						0.17	0.18	0.22±0.08
La Nina year	1981	1984	1988		1995	1998	1999			
T_{av}	23.28	25.30	23.72		24.00	24.07	24.51			24.15±0.70
S_{av}	33.89	34.06	34.01		33.60	34.18	33.90			33.94±0.20
DO_{av}	391.7	431.0	388.2		380.8	308.7	396.2			382.8±40.3
$\text{PO}_4\text{-P}_{av}$					0.20	0.23	0.56			0.33±0.20
$\text{SiO}_3\text{-Si}_{av}$					10.00	5.20	22.32			7.95±12.45
DIN_{av}					3.50	3.00	2.31			2.94±5.98
Chl- <i>a</i> ^b		0.08, 0.06				0.08	0.09			0.08±0.01

^a Mean±SD is the multi-year mean value ± standard deviation for the El Nino/La Nina years.

^b The Chl-*a* data were derived from SeaWiFS during 1998–2004, from Nimbus-7 CZCS in 1980 and 1984, provided by X. Chen, from The Third Institute of Oceanography, SOA, China for 1982, Fan (1985) and Huang (1992) for 1990, respectively.

Table 6. Comparison of the multi-year mean values of Chl-*a*, primary production (PP), phytoplankton abundance (PA), zooplankton biomass (ZB), benthos biomass (BB), cephalopod catch (CC) and demersal trawl catch (DTC) in the nSCS between two periods, i.e. before and after 1997 (units: mg m^{-3} , $\text{mg C m}^{-2} \text{d}^{-1}$, $\times 10^3 \text{ cell m}^{-3}$, mg m^{-3} , g m^{-2} , $\times 10^4 \text{ t}$, and $\times 10^4 \text{ t}$, respectively)^a.

Parameter	First phase (data source)	Second phase (data source)
Chl- <i>a</i>	0.29 (Fan 1985) ^b	2.16 (Cai et al., 2002a; SOAC 2003) ^c
PP	185.7 (Fan, 1985) ^b	442.5 (Ning et al., 2004; Jia et al., 2005; Hao et al., 2007)
PA	161.5 (Lin, 1985)	542.7 (Jia et al., 2005) ^b
ZB	44.5 (Zhang, 1984; Chen, 1985; Qian et al., 1990a, b; Huang et al., 1990; Chen, 1992)	22.05 (Jia et al., 2005) ^d
BB	6.6 (Shen, 1985)	11.3 (Jia et al., 2005)
CC	1.2 (Guo and Chen, 2000)	9.8 (Guo and Chen, 2000)
DTC	89.0 (Guo and Chen, 2000)	334.8 (Guo and Chen, 2000)

^a The data listed in the Table are the mean of the water column of the multi-year average for each parameter from different sources, except for BB, CC and DTC which are only treated by the multi-year average;

^b The data were provided by Chen, X. from The Third Institute of Oceanography, SOA, China;

^c The data was from the National Science Foundation of China (NSFC) project under the contract No. 90211021;

^d The data from Jia et al. (2005) are the annual mean of the multi-year mean for each season during 1997–2002.

same as above row, Jia et al., 2003). And nutrients were also low, i.e. the mean values of $\text{PO}_4\text{-P}=0.28 \mu\text{mol L}^{-1}$ and $\text{DIN}=3.63 \mu\text{mol L}^{-1}$ (according to criterion of Nutrition grade of sea water of China, Jia et al., 2003). The nutrients values (Jia et al., 2005, nutrients) are much lower than those we observed and the average index of ecological environment quality of fishing ground is 0.58, indicating it was in good condition (Jia et al., 2005).

The first phase (before 1997) represents the initial status of the biology and fisheries during the study period, and the second phase (after 1997) represents the response of the ecosystem to the changed environment, particularly since 1998, when the occurrence of N limitation was significantly alleviated. These ecosystem responses discussed above were clearly the result of environmental changes induced by not only climate change, but also anthropogenic activities.

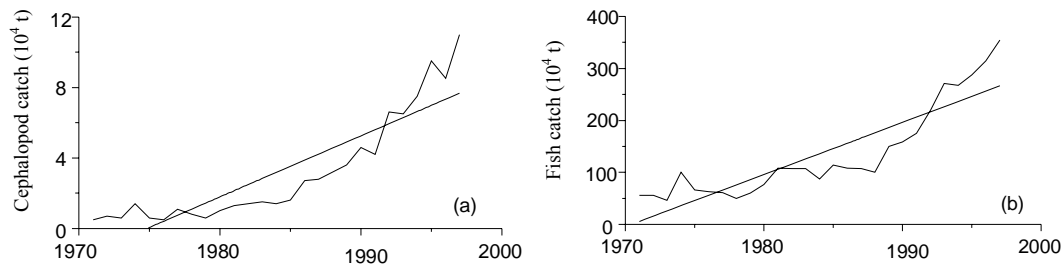


Fig. 12. Variation trends in the catches of cephalopod and fishes of the nSCS. (a) Catches of cephalopod; (b) Catches of fishes (Data from Guo and Chen, 2000).

5 Conclusion

The fluctuations in environmental parameters in the nSCS during 1976–2004 displayed different patterns, i.e. temperature (SST and T_{av}), salinity (S_{av} and S_{200}), dissolved inorganic nitrogen (DIN) and N:P annual rates increased, while DO, P, Si, Si:N, SSS and T_{200} annual rates decreased. The climate trend coefficients, R_{xt} of these time series were all over 0.38 ($n=29$) or 0.50 ($n=16$), and highly significant ($p \leq 0.05$), except for the time series of P and Si.

The increasing trends in SST and T_{av} were consistent with the rise of the mean air temperature (AT) in the Northern Hemisphere and southern China. The increase in SST and T_{av} and decrease in SSS in the nSCS led to strengthening of the thermocline and halocline, less mixing of deep water to the surface and thus a decrease in the P supply from deep waters. The increasing trend in DIN may have been influenced by the Pearl River discharge, and atmospheric dry and wet deposition, (which are related to anthropogenic activities), coastal upwelling and cyclonic eddies. The nSCS always experienced limitation of N before 1997, but the situation in the upper layer sea water has been mitigated since 1998, due to the increase in N concentration and decrease in P, which resulted in not only the positive trends in N:P ratios which are now close to the Redfield ratio, but also the decreasing trend in Si:N ratios, indicating of potential P limitation. The decrease in DO concentration may be linked to the increase in seawater temperature and the increase in the concentration of organic matter inputs mainly from the Pearl River and phytoplankton blooms, particularly since the 1990s. Chlorophyll-*a*, primary production, phytoplankton abundance, benthos biomass, cephalopod catch and demersal trawl catch have increased, and zooplankton abundance decreased. These ecosystem responses resulted from environmental changes were induced by not only climate change, but also anthropogenic activities. After 1998, phosphorus depletion in upper layer may be associated with a shift from diatoms to dinoflagellates and cyanophytes domination.

Pronounced responses of the environmental parameters to ENSO were observed. The effects of climate change on the nSCS were mainly through changes in monsoon and its

causative links, monsoon – circulation – nutrients – primary production.

It must be pointed out that although the data set we adopted in the present analysis is large and from various sources. As such the quality of data might vary throughout the long period of observation. But these data were valuable for studying the long term changes in environmental conditions and the responses of the ecosystem in this region. And using these data we can still find and understand some regulations on the response of the ecosystems to the environmental changes in the SCS. This regional response of SCS to global climate change was investigated the first which is far from well understood. The evolving nutrient environment may be related to the observed ecosystem changes in the nSCS such as increase in biological productivity. More long-time series observations on the structure and function of ecosystems and the relationships with environmental changes are needed in the SCS in the future.

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