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Concentration and vertical flux of Fukushima-derived radiocesium in sinking particles from two sites in the Northwestern Pacific Ocean

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Abstract. At two stations in the western North Pacific, K2 in the subarctic gyre and S1 in the subtropical gyre, time-series sediment traps were collecting sinking particles when the Fukushima Daiichi Nuclear Power Plant (FNPP1) accident occurred on 11 March 2011. Radiocesium (^{134}Cs and ^{137}Cs) derived from the FNPP1 accident was detected in sinking particles collected at 500 m in late March 2011 and at 4810 m in early April 2011 at both stations. The sinking velocity of ^{134}Cs and ^{137}Cs was estimated to be 22 to 71 m day^{-1} between the surface and 500 m and > 180 m day^{-1} between 500 m and 4810 m. ^{137}Cs concentrations varied from 0.14 to 0.25 Bq g^{-1} dry weight. These values are higher than those of surface seawater, suspended particles, and zooplankton collected in April 2011. Although the radiocesium may have been adsorbed onto or incorporated into clay minerals, correlations between ^{134}Cs and lithogenic material were not always significant; therefore, the form of the cesium associated with the sinking particles is still an open question. The total ^{137}Cs inventory by late June at K2 and by late July at S1 was 0.5 to 1.7 Bq m^{-2} at both depths. Compared with ^{137}Cs input from both stations by April 2011, estimated from the surface ^{137}Cs concentration and mixed-layer depth and by assuming that the observed ^{137}Cs flux was constant throughout the year, the estimated removal rate of ^{137}Cs from the upper layer (residence time in the upper layer) was 0.3 to 1.5 % yr^{-1} (68 to 312 yr). The estimated removal rates and residence times are comparable to previously reported values after the Chernobyl accident (removal rate: 0.2–1 %, residence time: 130–390 yr).

1 Introduction

On 11 March 2011, the 2011 Tohoku-Oki Earthquake of magnitude 9.0 occurred off Miyagi Prefecture, Japan. This earthquake and the tsunami it generated seriously damaged the Fukushima Daiichi Nuclear Power Plant (FNPP1). As a result of the loss of power and the associated malfunction of the cooling system, large quantities of artificial radionuclides were emitted from FNPP1 by hydrogen explosions, venting, and intentional and accidental discharge of contaminated water. However, the exact amount of radionuclides emitted from FNPP1 are still being argued: regarding eolian/atmospheric input, the estimates of ^{137}Cs emission range from about 5 to 35 PBq (e.g. 20–53 with average of 37 PBq, Stohl et al., 2011; 5.5–9.7 PBq, Miyazawa et al., 2013), and, regarding direct discharge to the ocean, estimates of ^{137}Cs discharge range from about 3 to 40 PBq (e.g. 12–24 with average of 27 PBq, Bailly du Bois, 2012; 3.5 ± 0.7 PBq, Tsumune et al., 2012; 11–16 PBq, Charette et al., 2013; 16.2 ± 1.6 PBq, Rypina et al., 2013). Based on estimates by Japanese scientists, an estimated 15 PBq of ^{137}Cs was emitted by hydrogen explosions and intentional venting (Chino et al., 2011; NERH, 2011). This amount corresponds to about one-sixth of the quantity released by the Chernobyl nuclear power plant (CNPP) accident in April 1986 (85 PBq; WHO, 1989). Morino et al. (2011) reported that only about 22 % of the ^{137}Cs released from FNPP1 was deposited on land in March 2011. Thus, one of the big differences between the FNPP1 accident and the CNPP accident is that, because FNPP1 is in a coastal area, most of the released radionuclides

(about 80 % of total emissions) were deposited in the ocean. Direct discharge of ^{137}Cs in contaminated water was estimated to be 3.5–5.9 PBq (Kawamura et al., 2011; Tsumune et al., 2011; Miyazawa et al., 2012). In addition, eolian input of ^{137}Cs was estimated to be 5 to 10 PBq (Kawamura et al., 2011; Miyazawa et al., 2012; Aoyama et al., 2012). About one month after the FNPP1 accident, Honda et al. (2012) investigated the dispersion of FNPP1-derived radiocesium in the western North Pacific. They measured radiocesium in surface water, suspended particles, and zooplankton and performed a mathematical simulation that verified that FNPP1-derived radiocesium had dispersed over a broad area of the western North Pacific. They suspected that radiocesium detected in remote parts of the North Pacific had been transported via the atmosphere and was not derived from the released contaminated water. Since then the dispersion of FNPP1-derived radionuclides has been investigated over an even broader area by voluntary observing ships and scientific cruises (Aoyama et al., 2012; Buesseler et al., 2012) and by mathematical simulations (Masumoto et al., 2012; Miyazawa et al., 2012). Nevertheless, the vertical invasion and transport of radionuclides in the ocean is still insufficiently understood. Radionuclides invade the ocean interior by vertical diffusion and advection or on sinking particles. After the CNPP accident, radionuclides on sinking particles were collected by sediment traps in the Black Sea (Buesseler et al. 1987, 1990), North Sea (Kempe and Nies, 1987), Mediterranean Sea (Fowler et al., 1987), and North Pacific (Kusakabe et al., 1988), and the arrival dates of the CNPP-derived radionuclides at the different stations as well as the sinking velocity and removal rate of the particles were determined.

At the time of the FNPP1 accident, sediment traps had already been deployed at time-series stations in the northwestern North Pacific to collect sinking particles for studies of biogeochemical cycling and how it would be affected by climate and oceanic change (project http://www.jamstec.go.jp/rigc/e/ebcrp/mbcrt/index.html). Here, we document the concentration and flux of FNPP1-derived radiocesium, ^{134}Cs (half life, 2.06 yr) and ^{137}Cs (half life, 30.17 yr), in sinking particles collected at these stations before and after the FNPP1 accident. Then, we estimate the sinking velocity of the particles and discuss the form and particle scavenging process of the radiocesium.

2 Method

2.1 Sediment trap positions

In November 2010, time-series sediment traps (McLane Mark7G-21) were deployed at 500 m and 4810 m at stations K2 (47° N, 160° E; water depth, ∼ 5200 m), which is in the subarctic gyre, and S1 (30° N, 145° E; water depth, ∼ 5800 m), which is in the subtropical gyre (Fig. 1). The horizontal distances from FNPP1 to K2 and S1 are about
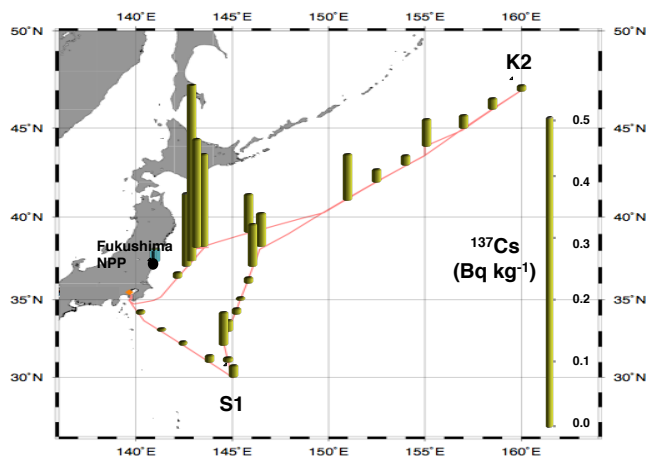

Fig. 1. Sediment trap location. Background is horizontal distribution of ^{137}Cs concentration in surface seawater along R/V *Mirai* MR11-02 cruise track in April 2011 (Honda et al., 2012).

1870 km and 950 km, respectively. Before the traps were deployed, their collecting cups were filled with a seawater-based 10 % buffered formalin solution as a preservative. The sampling interval was 12 days. The sediment traps at K2 and S1 were recovered in June and July 2011, respectively, by R/V *Mirai*.

2.2 Chemical analysis

In the laboratory, sediment trap samples were sieved through a 1 mm plastic mesh to eliminate zooplankton "swimmers" larger than 1 mm. The samples were then divided into 10 or more fractions by using a McLane splitter. Some fractions were filtered, rinsing with a small amount of distilled water, through pre-weighed Nuclepore filters (0.4 μm pore size). If swimmers were found in sample after filtration, these were removed by a tweezers. Thereafter samples were dried at 50 °C for 24 h. The dried samples were then weighed and total mass fluxes were computed. Subsequently, the samples were scraped off the filters and pulverized with an agate mortar.

For measurement of the total carbon concentration (TC), samples of about 5 mg were precisely weighed in a tin capsule. Samples for the measurement of organic carbon (OC) were precisely weighed on GF/F filters, and then the GF/F filters with sample were placed in a container filled with HCl mist for 24 h to eliminate CaCO$_3$. After decalcification, the filter with sample was wrapped in a tin disk. Samples in the tin capsule or wrapped in the tin disk were placed in the autosampler of an elemental analyzer (Perkin-Elmer 2400), and TC, OC and the concentration of nitrogen (N) were measured. The concentration of inorganic carbon (IC) was determined as the difference between TC and OC. Measurement errors were mainly less than 3 %.
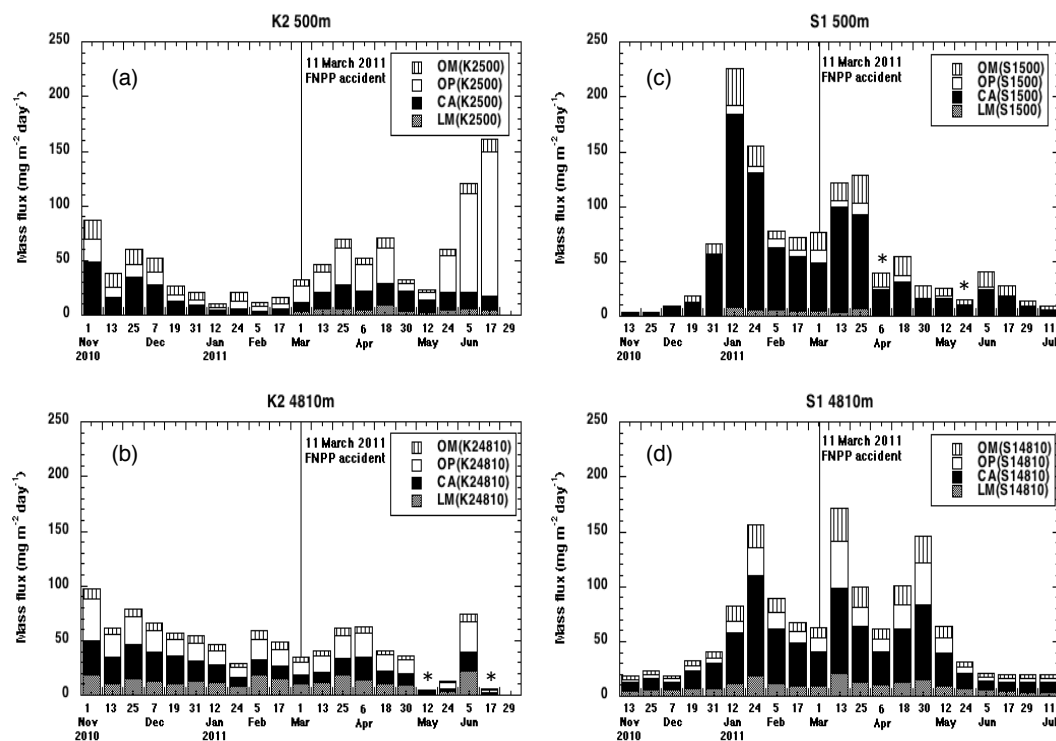
Fig. 2. Seasonal variability of sinking particles at **(a)** K2-500 m, **(b)** K2-4810 m, **(c)** S1-500 m and **(d)** S1-4810 m. Chemical compositions of OM, OP, CA and LM express organic materials, biogenic opal, CaCO$_3$ and lithogenic materials, respectively. Date is opening day of respective collecting cups. Radiocesium was measured on sinking particle samples collected after 1 March 2011. From samples with asterisk, radiocesium derived from the FNPP1 accident was not detected because of insufficient sample mass.

For measurement of trace elements, samples of about 20 mg were precisely weighed in a graphite crucible. About 140 mg of lithium metaborate (LiBO$_4$) (approximately seven times the sample mass) was then added and mixed well with the sample. Samples in graphite crucibles were placed in a furnace and maintained for 15 min at 950 °C. The resulting "lava-like" samples were placed in a Teflon beaker with 6 % nitric acid solution (HNO$_3$, ∼ 20 mL) and stirred for 1 h. This solution was filtered through a GF/F filter to eliminate graphite debris, its volume was adjusted to 50 ml by adding 6 % HNO$_3$ solution, and then it was weighed precisely. Concentrations of trace elements (Si, Ca, Fe, Ti, Mn, Mg, Ba, K) were measured by ICP-AES (Perkin-Elmer Optima 3300DV). Measurement errors were less than 5 % for every element. Concentrations of organic materials (OM), CaCO$_3$, biogenic opal and lithogenic material (LM) were calculated following Honda et al. (2002).

Radionuclide activities of ^{134}Cs and ^{137}Cs (hereinafter concentrations) were measured in samples collected after 1 March 2012. A plastic tube was filled with dried and pulverized samples, and the specific gamma rays emitted by ^{134}Cs and ^{137}Cs were measured by gamma spectrometry with a well-type Ge detector (ORTEC GWL-120210). The Ge detector was calibrated by using mixed-volume sources (Eckert & Ziegler Isotope Products, EG-ML). The counting time var-

ied from one day to a few days. The ^{134}Cs activity was corrected for cascade summing (43–49 %). The detection limit was 10 mBq per sample. Concentrations of radiocesium were decay-corrected to the date and time of the middle day of the sampling period. The precision of ^{134}Cs and ^{137}Cs measurements was one sigma of the counting error.

3 Results

3.1 Seasonal variability of the flux and chemical composition of sinking particles

3.1.1 K2

The total mass flux (TMF) of sinking particles at 500 m at K2 (K2-500 m) decreased from November 2010 to February 2011 (Fig. 2a). After the FNPP1 accident in March 2011, TMF increased and, although it dipped in early May, the highest value was reached in June. Between November and February, CaCO$_3$ was the dominant component of the samples, whereas biogenic opal was dominant after March. During the entire observation period, biogenic opal and CaCO$_3$ concentrations were estimated to average 47 % and 31 %, respectively. At K2-4810 m, the TMF was not as large as at 500 m, but its seasonal variability was similar; it decreased
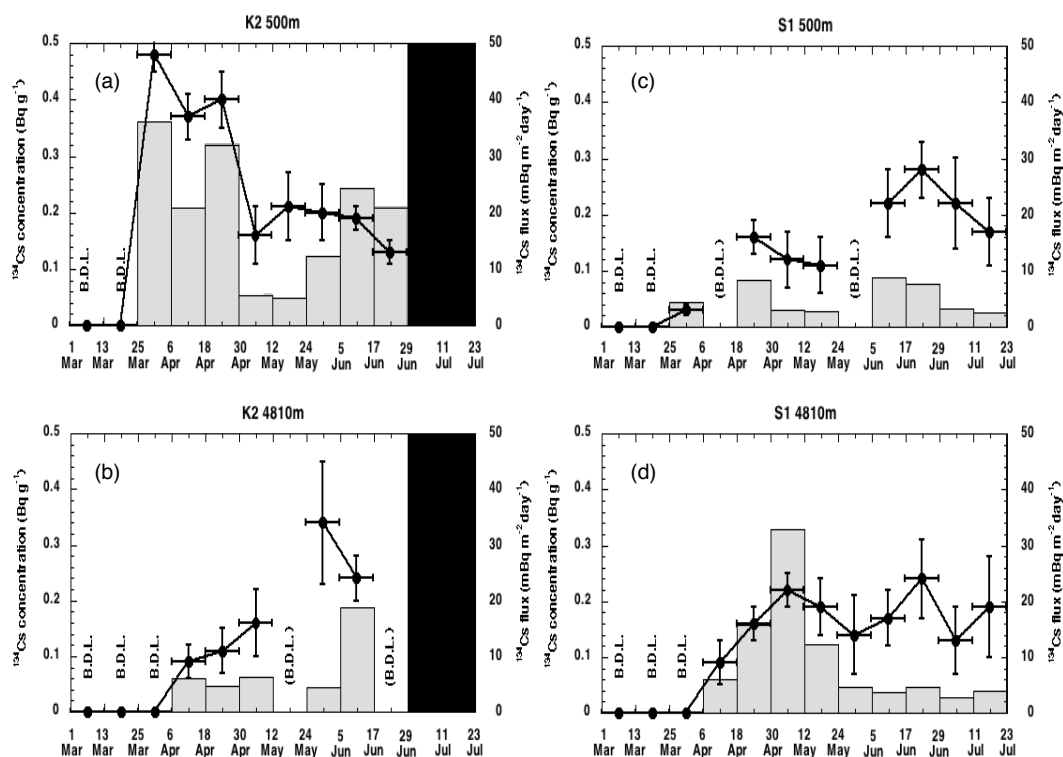
Fig. 3. Variability of ^{134}Cs concentration (line graph) and ^{134}Cs flux at **(a)** K2-500 m, **(b)** K2-4810 m, **(c)** S1-500 m and **(d)** S1-4810 m. B.D.L. is below detection limit. B.D.L. in parenthesis is supposed to be below detection limit likely caused by insufficient sample mass and radiocesium for gamma ray analysis.

from November 2010 and mainly increased, with low values in May, from March to June 2011 (Fig. 2b). The LM concentration was relatively higher at 4810 m (24 %) than at 500 m (6 %). We attribute this result to LM being more refractory than biogenic materials. Biogenic opal accounted for about 33 % of each sample, and was also dominant in the annual average. Therefore, the sediment trap samples form K2 used for the radiocesium analysis in this study consisted of sinking particles, of which biogenic opal was the major component, and the TMF tended to increase from March toward early summer.

3.1.2 S1

The smallest TMF was observed at S1-500 m at the time that the sediment trap was deployed in November 2010 (Fig. 2c). It began to increase in late December 2010 and reached a maximum in January 2011, after which it decreased before rising again in March 2011. Thereafter, TMF tended to decrease toward July 2011. The CaCO$_3$ concentration was about 80 % on average, and CaCO$_3$ was the major component of sinking particles. At S1-4810 m, TMF also increased from late December 2010 to January 2011 (Fig. 2d), and TMF peaks were observed in March and late April 2011. In general, the seasonal variability at 4810 m was similar to that at 500 m. Although concentrations of biogenic opal and LM

in the samples were higher at 4810 m than at 500 m, CaCO$_3$ was still the dominant component at 4810 m. Therefore, samples for radiocesium analysis from S1 consisted of sinking particles whose major component was CaCO$_3$, and the TMF tended to gradually decrease from March toward June.

3.2 Concentrations and flux of radiocesium

The concentrations and fluxes of ^{134}Cs and ^{137}Cs are listed in Table 1 and shown against time in Fig. 3. ^{134}Cs was detected in almost all samples, whereas before the FNPP1 accident, ^{134}Cs concentration was zero worldwide (e.g., Aoyama and Hirose, 2004). In addition, in all samples the activity ratio of ^{134}Cs to ^{137}Cs (^{134}Cs/^{137}Cs) was about one, which was the ratio in the radionuclides released by the FNPP1 accident, as verified by radiocesium analysis of seawater (Buesseler et al., 2011, 2012), aerosols (Haba et al., 2012) and soil (Qin et al., 2012). Thus, there is no doubt that the radiocesium detected in the sinking particles was derived from the FNPP1 accident.

3.2.1 K2

At K2-500 m, ^{134}Cs and ^{137}Cs were detected for the first time in sinking particles collected between 25 March and 6 April 2011 (Fig. 3a). The ^{134}Cs concentration in this sample, 0.48 Bq g^{-1}, was the maximum among

Table 1. Concentrations and fluxes of radiocesium at K2. The analytical error is based on one sigma of counting statistics. Concentrations are decay-corrected based on middle day of each sampling period. B.D.L. is below detection limit (0.01 Bq). (B.D.L.) is supposed to be below detection limit because of insufficient sample mass and activity by gamma procedures.

Station Depth	Opening Day	^{134}Cs concentration (Bq g^{-1})	^{137}Cs concentration (Bq g^{-1})	^{134}Cs/^{137}Cs	TMF (mg m^{-2} day^{-1})	^{134}Cs Flux (mBq m^{-2} day^{-1})	^{137}Cs Flux (mBq m^{-2} day^{-1})
K2 500 m	1-Mar	B.D.L.	B.D.L.	–	35.0	–	–
	13-Mar	B.D.L.	B.D.L.	–	51.2	–	–
	25-Mar	0.48 ± 0.03	0.41 ± 0.01	1.16 ± 0.07	75.4	36.2 ± 2.1	31.1 ± 0.6
	6-Apr	0.37 ± 0.04	0.35 ± 0.01	1.06 ± 0.13	55.9	20.9 ± 2.4	19.8 ± 0.8
	18-Apr	0.40 ± 0.05	0.34 ± 0.02	1.16 ± 0.15	80.0	31.9 ± 3.8	27.5 ± 1.5
	30-Apr	0.16 ± 0.05	0.16 ± 0.02	1.00 ± 0.33	33.6	5.3 ± 1.7	5.4 ± 0.6
	12-May	0.21 ± 0.06	0.19 ± 0.02	1.11 ± 0.36	23.9	4.9 ± 1.5	4.4 ± 0.6
	24-May	0.20 ± 0.05	0.20 ± 0.02	0.98 ± 0.24	61.5	12.4 ± 2.9	12.6 ± 1.0
	5-Jun	0.19 ± 0.02	0.19 ± 0.01	1.00 ± 0.14	125.6	24.3 ± 3.1	24.3 ± 1.2
	17-Jun	0.13 ± 0.02	0.12 ± 0.01	1.07 ± 0.22	166.9	20.9 ± 4.1	19.6 ± 1.4
K2 4810 m	1-Mar	B.D.L.	B.D.L.	–	37.5	–	–
	13-Mar	B.D.L.	B.D.L.	–	43.2	–	–
	25-Mar	B.D.L.	B.D.L.	–	66.4	–	–
	6-Apr	0.09 ± 0.03	0.07 ± 0.01	1.16 ± 0.49	69.6	6.1 ± 2.3	5.2 ± 0.9
	18-Apr	0.11 ± 0.04	0.11 ± 0.01	0.97 ± 0.35	43.3	4.7 ± 1.6	4.9 ± 0.6
	30-Apr	0.16 ± 0.06	0.17 ± 0.02	0.94 ± 0.39	39.9	6.3 ± 2.5	6.6 ± 0.9
	12-May	(B.D.L.)	(B.D.L.)	–	5.0	–	–
	24-May	0.34 ± 0.11	0.33 ± 0.04	1.01 ± 0.37	13.3	4.5 ± 1.5	4.4 ± 0.6
	5-Jun	0.24 ± 0.04	0.22 ± 0.02	1.10 ± 0.21	79.3	18.8 ± 3.1	17.1 ± 1.5
	17-Jun	(B.D.L.)	(B.D.L.)	–	5.8	–	–

all sediment trap samples analyzed in this study. Subsequently, ^{134}Cs concentration tended to decrease gradually. The largest ^{134}Cs flux was also observed between 25 March and 6 April (36.2 mBq m^{-2} day^{-1}), and it also tended to decrease thereafter.

At K2-4810 m, ^{134}Cs was initially detected in sinking particles collected during the next sampling interval, between 6 and 18 April, 12 days after it was first detected at K2-500 m (Fig. 3b). Unlike at K2-500 m, at K2-4810 m ^{134}Cs concentration increased over time, and it reached its largest value of 0.35 Bq g^{-1} between 24 May and 6 June 2011. No ^{134}Cs was detected in sinking particles collected between 12 and 24 May or between 17 and 29 June, probably because the sample mass was insufficient for detection by the gamma procedure rather than because the sinking particles contained no FNPP1-derived ^{134}Cs. At K2-4810 m, the ^{134}Cs flux was largest (20.9 mBq m^{-2} day^{-1}) between 5 and 17 June.

The average ^{134}Cs concentration at K2-500 m and K2-4810 m during the observation period (500 m, 96 days; 4810 m, 60 days) was 0.27 and 0.19 Bq g^{-1}, respectively. The total ^{134}Cs inventory was 1.9 Bq m^{-2} at 500 m and 0.5 Bq m^{-2} at 4810 m (Table 2).

3.2.2 S1

At S1-500 m, ^{134}Cs was first detected in sinking particles collected between 25 March and 6 April, that is, the same period that they were first detected at K2-500 m (Fig. 3c). In general, ^{134}Cs concentration tended to increase with time at S1-500 m, and it reached a maximum ($\sim$ 0.3 Bq g^{-1}) in sinking

particles collected between 17 and 29 June. As at K2-4810 m, and probably for the same reason, no ^{134}Cs was detected in sinking particle samples collected at S1-500 m between 6 and 18 April and between 24 May and 6 June. The ^{134}Cs flux was relatively higher in late April and June, and variations of the flux and the activity were generally synchronous. The total ^{134}Cs inventory and average ^{134}Cs activity during the observation period (96 days) were estimated to be 0.5 Bq m^{-2} and 0.16 Bq g^{-1}, respectively (Table 2).

At S1-4810 m ^{134}Cs was initially detected in sinking particles collected between 6 and 18 April (Fig. 3d), the same period that ^{134}Cs was first detected at K2-4810 m (Fig. 3b). ^{134}Cs concentration increased toward late April and early May, and thereafter it fluctuated around a relatively constant value. A maximum ^{134}Cs concentration of 0.25 Bq g^{-1} was detected in the sample collected between 17 and 29 June. The peak ^{134}Cs flux occurred between 30 April and 12 May, after which it decreased. The total ^{134}Cs flux and the average ^{134}Cs activity at S1-4810 m were estimated to be 1.0 Bq m^{-2} and 0.17 Bq g^{-1}, respectively (Table 2).

Previous ocean observations and mathematical simulation results suggest that the radiocesium detected at K2 and S1 in this study likely did not derive from contaminated water discharged from FNPP1, nor subsequently transported to the sediment trap stations, but from contaminated eolian dust (Honda et al., 2012).

Table 2. Concentrations and fluxes of radiocesium at S1. The analytical error is based on one sigma of counting statistics. Activities are decay-corrected based on the middle day of each sampling period. B.D.L. is below detection limit (0.01 Bq). (B.D.L.) is supposed to be below detection limit because of insufficient sample mass and activity by gamma procedures.

Station Depth	Opening Day	^{134}Cs concentration (Bq g^{-1})	^{137}Cs concentration (Bq g^{-1})	^{134}Cs/^{137}Cs	TMF (mg m^{-2} day^{-1})	^{134}Cs Flux (mBq m^{-2} day^{-1})	^{137}Cs Flux (mBq m^{-2} day^{-1})
S1 500 m	1-Mar	B.D.L.	B.D.L.	–	77.9	–	–
	13-Mar	B.D.L.	B.D.L.	–	127.4	–	–
	25-Mar	0.03 ± 0.01	0.03 ± 0.00	1.24 ± 0.44	126.3	4.4 ± 1.5	3.5 ± 0.5
	6-Apr	(B.D.L.)	(B.D.L.)	–	37.2	–	–
	18-Apr	0.16 ± 0.03	0.16 ± 0.01	0.98 ± 0.20	52.3	8.3 ± 1.6	8.5 ± 0.6
	30-Apr	0.12 ± 0.05	0.11 ± 0.03	1.16 ± 0.57	24.9	3.1 ± 1.3	2.6 ± 0.6
	12-May	0.11 ± 0.05	0.10 ± 0.02	1.06 ± 0.53	26.1	2.8 ± 1.3	2.7 ± 0.5
	24-May	(B.D.L.)	(B.D.L.)	–	16.6	–	–
	5-Jun	0.22 ± 0.06	0.23 ± 0.02	0.95 ± 0.27	40.5	8.9 ± 2.3	9.3 ± 0.9
	17-Jun	0.28 ± 0.05	0.21 ± 0.02	1.36 ± 0.27	27.6	7.8 ± 1.4	5.7 ± 0.5
	29-Jun	0.22 ± 0.08	0.20 ± 0.03	1.10 ± 0.43	14.6	3.2 ± 1.2	2.9 ± 0.4
	11-Jul	0.17 ± 0.06	0.18 ± 0.03	0.92 ± 0.38	15.1	2.5 ± 1.0	2.7 ± 0.4
S1 4810 m	1-Mar	B.D.L.	B.D.L.	–	63.0	–	–
	13-Mar	B.D.L.	B.D.L.	–	170.4	–	–
	25-Mar	B.D.L.	B.D.L.	–	98.2	–	–
	6-Apr	0.09 ± 0.04	0.07 ± 0.01	1.31 ± 0.57	63.8	6.0 ± 2.3	4.5 ± 0.9
	18-Apr	0.16 ± 0.03	0.13 ± 0.01	1.25 ± 0.25	100.0	15.7 ± 2.9	12.6 ± 1.2
	30-Apr	0.22 ± 0.03	0.19 ± 0.01	1.20 ± 0.18	147.0	32.9 ± 4.4	27.3 ± 1.7
	12-May	0.19 ± 0.05	0.18 ± 0.02	1.04 ± 0.26	64.6	12.3 ± 2.9	11.9 ± 1.1
	24-May	0.14 ± 0.07	0.13 ± 0.03	1.08 ± 0.53	31.9	4.6 ± 2.1	4.3 ± 0.8
	5-Jun	0.17 ± 0.05	0.12 ± 0.02	1.37 ± 0.50	21.7	3.6 ± 1.2	2.6 ± 0.5
	17-Jun	0.24 ± 0.07	0.18 ± 0.03	1.33 ± 0.42	19.2	4.6 ± 1.3	3.4 ± 0.5
	29-Jun	0.13 ± 0.06	0.11 ± 0.02	1.19 ± 0.63	20.8	2.7 ± 1.3	2.3 ± 0.4
	11-Jul	0.19 ± 0.09	0.16 ± 0.04	1.16 ± 0.66	20.9	3.9 ± 2.0	3.3 ± 0.9

4 Discussions

4.1 Sinking velocity

Although the distance to K2 from FNPP1 is double that to S1, it is noteworthy that at each respective depth FNPP1-derived radiocesium in sinking particles was first detected at both stations in the same period: at 500 m in samples collected between 25 March and 6 April, and at 4810 m in samples collected between 6 and 18 April. We estimated the sinking velocity of radiocesium between 500 m and 4810 m to be greater than ~ 180 m day^{-1} [(4810–500)/(12 + 12)]. This velocity is comparable to that estimated for sediment trap radionuclide data from the CNPP accident (~ 190 m day^{-1}: Kusakabe et al., 1988) and for general deep-sea sediment trap data (~ 140 m day^{-1}, Honda et al., 2009; 100–300 m day^{-1}, Berelson, 2002).

The results of a mathematical simulation of the transport of contaminated eolian dust containing radiocesium (Honda et al., 2012) suggest that eolian radiocesium arrived and was deposited at K2 on 14 March and at S1 on 18 March 2011. Under the assumption that the eolian radiocesium arrived at 500 m between 25 March and 6 April, we estimated the sinking velocity between 0 (surface) and 500 m as 22 to 46 m day^{-1} at K2 and 26 to 71 m day^{-1} at S1. Although these values are smaller than the estimated general deep-sea sinking

velocity of ~ 140 m day^{-1}, these low velocities at shallow depth are comparable to the estimated radionuclide flux at shallow depth associated with the CNPP accident (> 20 m day^{-1} in the upper 1000 m of the Black Sea, Buesseler et al., 1987; 29 m day^{-1} in the upper 200 m of the Mediterranean Sea, Fowler et al., 1987; 65 ± 22 m day^{-1} in the upper 222 m of the North Sea, Kempe and Nies, 1987). This increase in the sinking velocity with depth is also supported by previous work (Berelson, 2002; Honda et al., 2009). This increase might be partly attributed to the increase in density of settling particle because "light" and fragile organic material decreases with depth compared to "heavy" and relatively refractory opal, CaCO$_3$ and lithogenic material.

At K2-500 m, the sinking particles collected between 25 March and 18 April were composed of biogenic opal (~ 46 %) and CaCO$_3$ (~ 32 %), whereas at S1-500 m, they consisted of ~ 9 % biogenic opal and ~ 72 % CaCO$_3$. Thus, at the two stations, the sinking velocities were comparable even though the chemical composition of the sinking particles differed.

Table 3. Total radiocesium flux, average radiocesium concentration, suspected ^{137}Cs inventory (see text), and removal rate of ^{137}Cs from and residence time of ^{137}Cs at surface water.

	^{134}Cs concentration (Bq g^{-1})	^{134}Cs concentration (Bq g^{-1})	^{134}Cs Inventory (Bq m^{-2})	^{134}Cs Inventory (Bq m^{-2})	Sampling period (day)	^{137}Cs Flux (mBq m^{-2} day^{-1})	^{137}Cs inventory (Bq m^{-2})	Removal rate (% yr^{-1})	Residence time (yr)
K2 500	0.27 ± 0.13	0.25 ± 0.11	1.9	1.7	96	18.1	450	1.5	68
K2 4810	0.19 ± 0.10	0.18 ± 0.10	0.5	0.5	60	7.6	450	0.6	161
S1 500	0.16 ± 0.08	0.15 ± 0.07	0.5	0.5	96	4.7	540	0.3	312
S1 4810	0.17 ± 0.05	0.14 ± 0.04	1.0	0.9	108	8.0	540	0.5	184

4.2 Comparisons with radiocesium activities of sinking particles from the CNPP accident and those of seawater, suspended substances, and zooplankton

The ^{137}Cs concentration of sinking particles observed in this study ranged from 0.03 to 0.41 Bq g^{-1} (30–410 Bq kg^{-1}; Table 1). We estimated the average ^{137}Cs concentration to be 0.18 to 0.25 Bq g^{-1} (180–250 Bq kg^{-1}) at K2 and 0.14 to 0.15 Bq kg^{-1} (140–150 Bq kg^{-1}) at S1 (Table 2). Compared with the ^{137}Cs concentrations resulting from the FNPP1 accident observed in this study, the ^{137}Cs concentrations of sinking particles observed just after the CNPP accident were higher (Black Sea, 0.5 and 1.9 Bq g^{-1}, Buesseler et al., 1987; Mediterranean Sea, 0.15–4.0 Bq g^{-1}, Fowler et al., 1987; North Sea, 1.6–13.6 Bq g^{-1}, Kempe and Nies, 1987). Despite distances from the CNPP to respective sediment trap stations (about 580 km for the Black Sea, 1950 km for the Mediterranean Sea and 1880 km for the North Sea) compared to distances from the FNPP1 to S1 (950 km) and K2 (1870 km), concentrations observed after the CNPP accident were much higher than that after the FNPP1 accident this study. This might be mainly attributable to the fact that the amount of ^{137}Cs emitted during the CNPP accident was about six times the amount emitted during the FNPP1 accident.

In addition, as a result of wind direction and dry/wet deposition, the higher concentrations after the CNPP accident might be attributed to higher fallout of radionuclides to observation areas and subsequent elevated concentration in the water.

After the CNPP accident, the observed ^{137}Cs concentrations of surface seawater (Black Sea, 0.04–0.19 Bq kg^{-1}, Buesseler et al., 1987; North Sea, 0.07–0.3 Bq kg^{-1}, Kempe and Nies, 1987) were also higher than activities of surface seawater observed in April 2011 at K2 and S1 (0.009 and 0.018 Bq kg^{-1}, respectively), after the FNPP1 accident (Honda et al., 2012).

On the other hand, ^{137}Cs concentrations observed in sediment trap samples from the northern North Pacific and the Bering Sea after the CNPP accident (0.01–0.027 Bq g^{-1} in the North Pacific and 0.07–0.25 Bq g^{-1} in the Bering Sea, Kusakabe et al., 1987) were comparable to or smaller than those observed after the FNPP1 accident. These similar activities presumably reflect the fact that the distance between

the CNPP and the North Pacific and Bering Sea sediment trap stations is about 8500 km, so the amount of ^{137}Cs transported from the CNPP to these areas was small.

In contrast to the relatively low ^{137}Cs concentrations of surface seawater at K2 and S1 in April 2011, about one month after the FNPP1 accident (Honda et al., 2012), ^{137}Cs concentrations of suspended particles in April 2011 at both stations were quite high, ranging from 3 to 32 Bq kg^{-1}, and concentrations of zooplankton collected at the same time ranged from 13 to 71 Bq kg^{-1} dry weight, slightly higher than those of suspended particles (Honda et al., 2012). These ^{137}Cs concentrations of zooplankton are comparable to both those of zooplankton observed after the CNPP accident in the Mediterranean Sea (Fowler et al., 1987) and those observed in the western North Pacific after the FNPP1 accident (Buesseler et al., 2012). However, ^{137}Cs activities of sinking particles in this study (140–250 Bq kg^{-1} on average) were several times higher than those of zooplankton. In steady state, the distribution coefficient (K_ds) of Cs of sediment or lithogenic materials is much higher than the concentration factor (CF) of creature (IAEA, 2004). Although the observation periods were not steady state, and only transient ^{137}Cs concentrations of the various materials were measured, it is suspected that radiocesium in sinking particles was mainly associated with lithogenic materials.

4.3 Form of the radiocesium associated with sinking particles

Cesium, including radiocesium, is an alkaline metal that is regarded as soluble in water. The Chernobyl fallout was rapidly (within days) solubilized in seawater (Whitehead et al., 1988). However, a part of the cesium can be adsorbed onto or incorporated into aluminosilicates (clay minerals), similar to potassium (Qin et al., 2012, and references therein), and the adsorption or incorporation of cesium is not easily reversed (e.g., Kogure et al., 2012; Otosaka and Kobayashi, 2012). To determine the form of the radiocesium associated with sinking particles, we computed correlations between the concentration and flux of radiocesium with those of organic carbon, biogenic opal, CaCO$_3$, lithogenic materials (LM) and trace elements. We expected that radiocesium would correlate with LM or potassium, but the correlation

between the ^{134}Cs flux and potassium or LM fluxes was not always significant: correlation coefficients (r^2) between fluxes of Cs and LM (potassium) for respective depths of stations were 0.42 (0.37) on average. A possible reason for the failure to find a significant correlation is that the radiocesium concentrations and fluxes were transient values; consequently, the magnitude of radiocesium adsorption or incorporation might not be constant. Although it is suspected that radiocesium in sinking particles was mainly associated with lithogenic materials as described before, another possible reason is that the cesium was partly assimilated by living organisms (e.g. Kasamatsu and Ishikawa, 1997). Fowler et al. (1987) also reported that ^{137}Cs can become concentrated in fecal pellets. Moreover, Kaneyasu et al. (2012) reported that sulfate aerosol is a potential transport medium of radiocesium rather than lithogenic materials. In this case, sulfate Cs is easily dissolved in surface and then might be absorbed by creature. Any of these circumstances could cause the correlation between radiocesium and potassium or LM to not be significant. Thus, the form of the radiocesium associated with sinking particles is still an open question. Precise measurements of stable Cs (^{133}Cs), electron microscopic analyses, and characterization/extraction experiments are needed to resolve this question.

5 Removal rate and residence time

In April 2011, ^{137}Cs concentration of surface water was 0.009 Bq kg^{-1} at K2 and 0.018 Bq kg^{-1} at S1 (Honda et al., 2012), but Honda et al. (2012) did not detect radiocesium from FNPP1 in subsurface water (200 m). In April 2011, the surface mixed layer thickness (SML) was 100 m at K2 and 60 m at S1. If it can be assumed that ^{137}Cs decreased linearly from the surface to the bottom of the SML, the ^{137}Cs inventory in April 2011 can be estimated as 450 Bq m^{-2} at K2 and 540 Bq m^{-2} at S1 (Table 2). The average ^{137}Cs flux (^{137}Cs inventory/the observation period) to the ocean interior ranged from 7.6 to 18.1 mBq m^{-2} day^{-1} at K2 and from 4.7 to 8.0 mBq m^{-2} day^{-1} at S1 (Table 2). After this study, ^{137}Cs flux might decrease with time. However concentration of ^{137}Cs at K2-4810 m, S1-500 m and S1-4810 m did not begin to decrease. In addition, it is suspected that total mass flux at K2 might increase in summer. If these ^{137}Cs fluxes remained constant for a year, the removal rate at K2 can be estimated as 0.62 % yr^{-1} to 1.47 % yr^{-1} and the residence time as 68 to 161 years, whereas the estimated removal rate at S1 is 0.32 % yr^{-1} to 0.54 % yr^{-1} and the estimated residence time is 184 to 312 years. From sediment trap and seawater ^{137}Cs data collected in the Black Sea in the several months following the CNPP accident, Buesseler et al. (1987) estimated a removal rate and residence time of radiocesium of 0.77 % yr^{-1} and 130 yr, respectively. Kusakabe et al. (1988) reported removal rates of 1 % in the North Pacific and Bering Sea for radiocesium from the CNPP accident, and Fowler et

al. (1987) reported a removal rate of 0.2 % in the Mediterranean Sea. All of these removal rate estimates for radiocesium from the CNPP accident are comparable to our estimates for radiocesium from the FNPP1 accident. However, the ^{137}Cs concentration and flux might change with time, which might alter estimates of the removal rate and residence time. For example, using sediment trap data from an entire year, Buesseler et al. (1990) estimated a longer residence time (230–390 yr) in the Black Sea. Residence time estimated this study (68–312 yr) might be an underestimate.

6 Concluding remarks and future work

Sediment trap data revealed that FNPP1-derived radiocesium was quickly transported to the deep sea in the western North Pacific. Estimates of the removal rate and residence time were comparable to previous estimates for CNPP-derived radiocesium. Moreover, data from sinking particles collected at K2 and S1 after the end of the observation period of this study confirm the increasing trend of TMF at K2 and the decreasing trend at S1 (Honda et al., in preparation). In addition, except at K2-500 m, ^{134}Cs concentration of sinking particles is continuing to increase with time. These new data will reveal more details about the radiocesium flux associated with sinking particles. We also collected seafloor sediments at K2 and S1 in June/July 2011. The removal rate of radiocesium from the water column to the seafloor sediment and, conversely, its remobilization from sediments are also of considerable interest (e.g. Otosaka and Kobayashi, 2012). In addition, in July 2011 we deployed time-series sediment traps about 90 km off Fukushima, and data from these sediment traps will reveal not only the vertical transport of radiocesium to the ocean interior but also its lateral transport from the land or from shallower seafloor sediments.

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