

A modeling analysis of a heavy air pollution episode occurred in Beijing

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Abstract. The concentrations of fine particulate matter (PM) and ozone in Beijing often exceed healthful levels in recent years, therefore China is taking steps to improve Beijing's air quality for the 2008 Olympic Games. In this paper, the Models-3 Community Multiscale Air Quality (CMAQ) Modeling System was used to investigate a heavy air pollution episode in Beijing during 3–7 April 2005 to obtain the basic information of how heavy air pollution formed and the contributions of local sources and surround emissions. The modeling domain covered from East Asia with four nested grids with 81 to 3 km horizontal resolution focusing on urban Beijing. This was coupled with a regional emissions inventory with a 10 km resolution and a local 1 km Beijing emissions database. The trend of predicted concentrations of various pollutants agreed reasonably well with the observations and captured the main features of this heavy pollution episode. The simulated column concentration distribution of PM was correlated well with the MODIS remote sensing products. Control runs with and without Beijing emissions were conducted to quantify the contributions of non-Beijing sources (NBS) to the Beijing local air pollution. The contributions of NBS to each species differed spatially and temporally with the order of $PM_{2.5} > PM_{10} > SO_2 > \text{soil}$ for this episode. The percentage contribution of NBS to fine particle ($PM_{2.5}$) in Beijing was averaged about 39%, up to 53% at the northwest of urban Beijing and only 15% at southwest. The spatial distribution of NBS contributions for PM_{10} was similar to that for $PM_{2.5}$, with a slightly less average percentage of about 30%. The average NBS contributions for SO_2 and soil (diameter between 2.5 μm and 10 μm) were 18% and 10%. In addition, the pollutant transport flux was calculated and compared at different levels to investigate transport pathway and magnitude. It was found that the NBS contribution correlated with the transport flux, contributing 60% of PM_{10} concen-

tration in Beijing at the time of transport flux peak during a strong episode with a transport path from southwest to northeast.

1 Introduction

Regional air pollution over Mega-city clusters has been accompanied with the quick economical expansion in China. Mega-cities are significant sources of aerosols and oxidants, which are important atmospheric components having direct impacts on air quality and climate on a regional and even global scale. It is critical to understand the key chemical and physical processes controlling the concentrations of these species in these polluted regions. As a research hotspot, one of the dominant mega-city clusters is located in Beijing, the capital of China, and its surroundings areas (Fig. 1). With a population more than 14 million, Beijing's economy growth has been rapidly growing with an annual increasing rate exceeding 10% continuously. For example, the total energy consumption is increased from 38.5 million ton standard coal in 1999 to 45.1 million-ton standard coal in 2002, resulting in a very serious air pollution problem. After Beijing succeeded in the competition for the bid to host the 2008 Olympic Games, improving the air quality in Beijing has been one of the most important tasks for the Chinese government. After a series of air antipollution measures including increasing the natural gas supply, centralizing the heat supply, removing electric heat devices, and using of low sulfur coals, air quality in Beijing has improved gradually since 1997. The number of days of meeting Air Pollution Standard II or better increased from 56 days in 1998 to 254 days in 2005. However, the average concentration of main contaminants remained high in the past three years, indicating that emissions controls only in Beijing are not enough to solve the regional pollution problems over Beijing.

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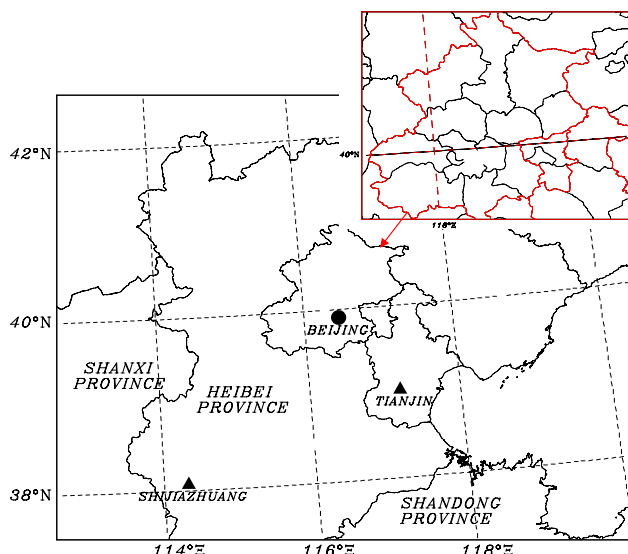


Fig. 1. Map of Beijing and surrounding areas.

Several studies pointed out that the regional air pollution in Beijing was caused not only by local sources but also significantly by non-Beijing sources (Zhang et al., 2004; Xu et al., 2004; Sun et al., 2004). The basin-like geophysical feature of urban Beijing and its unique weather system are the dominant factors influencing air quality (Ren et al., 2004). Su et al. (2004) investigated certain transport pathways of NBS to Beijing and found that a convergence zone of air pollutants occurred within the atmospheric boundary layer along the combination of the Yanshan-Taihang mountains and the North China Plain. The contribution of various emission sources shows a specific structure in spatial distribution and depends on the location and weather system.

In order to improve air quality in Beijing with comprehensive measures, it is necessary to quantitatively evaluate the contribution of NBS so that regional emissions can be controlled more efficiently. This can be achieved through a regional air quality modeling system. Yan et al. (2005) made a long-term simulation of SO_2 and discussed the impacts of different types of sources on air quality. The transport of pollutants among several cities in North China was investigated by Zhang et al. (2004) with simplified chemical transport model, which considered no chemical reactions in simulating PM_{10} .

In this study the formation process of a typical heavy air pollution episode in Beijing during 3–7 April 2005 was investigated with the Models-3/CMAQ. This episode clearly showed the typical evolution process of high particle pollution over Beijing with continuous several days' pollution. The temporal and spatial characteristics of the main contaminants were analyzed with a focus on the contributions of NBS to SO_2 , PM_{10} and 15 $\text{PM}_{2.5}$ concentrations in Beijing. The modeling results were also compared with observations of

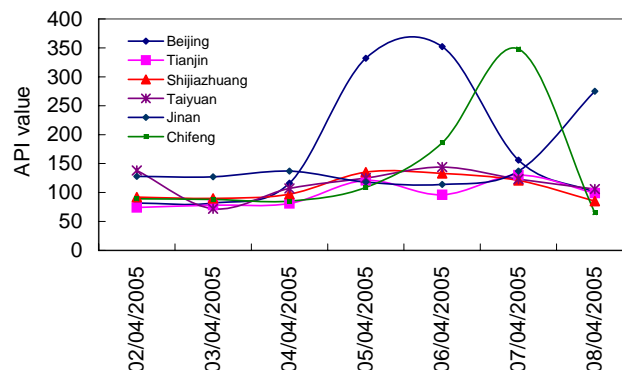


Fig. 2. API values of Beijing and surrounding cities from 2 to 8 April 2005.

SO_2 and particle matters at a 325 m meteorological tower located on the campus of the Institute of Atmospheric Physics (IAP), meteorological data observed at six monitoring sites in Beijing, and satellite aerosol remote sensing results from NASA.

2 Observations and regional patterns of the high air pollution episode

2.1 Air Pollution Index (API) of Beijing and surrounding cities

To indicate the air pollution level over the Chinese cities, an Air Pollutant Index (API) is issued every day by SEPA (State Environment Protection Agency of China) with the monitoring data of suspended particulate matter (PM_{10}), sulfur dioxide (SO_2), and nitrogen dioxide (NO_2) (<http://www.sepa.gov.cn>). An index value under 100 indicates the air quality of that monitoring station conforms to the National ambient air quality criteria II. A value over 100 will worsen the allergy symptom of unhealthy people. Figure 2 shows the time series of daily API of Beijing and surrounding cities in 2–8 April 2005. An obvious high pollution episode occurred in Beijing with API over 100 lasting more than 4 days from 4–7 April and reaching heavy pollution levels with a maximum API of 350 on 5–6 April.

2.2 SO_2 and particles observed at IAP 325 m tower

Six platforms were made on the IAP tower for simultaneous collection of samples for analysis of SO_2 and particulate matter. The 43CTL instruments of high precision impulse fluorescence (USA Thermion Environment Corporation) were used for measuring gas at 8 m, 47 m, 120 m, and 280 m. The RP1400a instruments (USA ANPU) were used for measuring mass concentration of the particles at 8 m and 240 m. The available data shown in Fig. 3 for this study include SO_2 at 47 m and PM_{10} at 8 m. From Fig. 3, it can be seen that SO_2

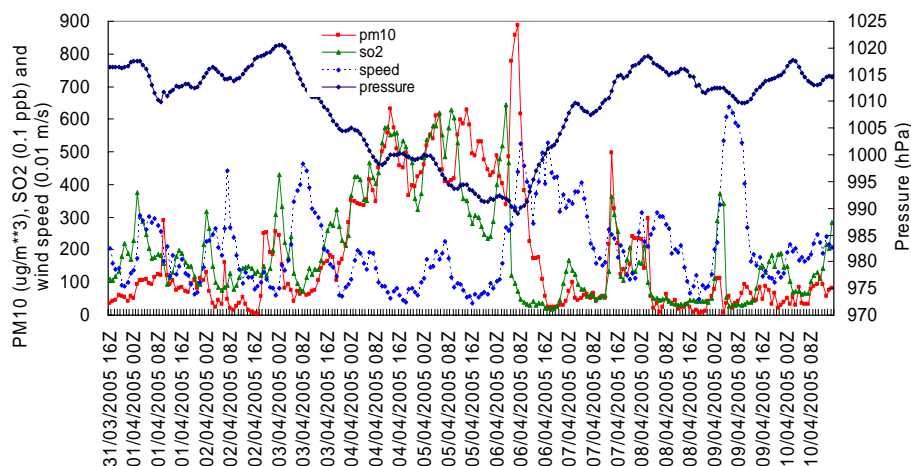


Fig. 3. Temporal variations of observed SO₂ and particle concentrations of IAP site and average pressure, wind speed of six observations sites in Beijing from 31 March 2005 16:00 to 10 April 2005 16:00 (UTC).

concentrations were high on 4–6 April, with peak value exceeding 60 ppb. Hourly variations of PM₁₀ has similar patterns with SO₂ (Fig. 3), but with much higher levels reaching 300 $\mu\text{g}/\text{m}^3$ for PM₁₀ during 4–6 April, respectively.

2.3 Meteorological patterns and controlling factors

Through the analysis of synoptic meteorological charts for this period, it was found that initial weak winds and stable weather patterns prevailed and later followed by transport of Asian dust from Inner Mongolia on 5 April causing the 4-day high of air pollution mixing of anthropogenic pollutants and natural dust particles. Surface synoptic patterns show that from 4–6 April, there is a strong high pressure system located over Yangtze River with the low in Beijing. This typical kind of surface pattern is unfavorable for pollution in Beijing to diffuse and also it is a typical type of weather pattern causing Beijing's serious pollution. Detailed analyzing the synoptic chart during this period, on 500 hPa chart, at 00:00 Z on 4 April of 2005, Beijing was in strong northwest flow at the back of the East Asia groove. On surface chart, there was a strong high-pressure center located at the southeast of China (Yangtze area), and Beijing lay at the top of the high-pressure. From 500 hPa chart at 00:00 Z on 5 April, we could see Beijing was in southwest flow at the front of a groove. Corresponding to 850 hPa and surface of the same time, the high-pressure center still existed, and from surface to high altitude, wind direction of Beijing was southwest. When it came to 06:00 Z, on 5 April, there was a Mongolia cyclone on 500 hPa altitude, and on surface, there were cold front, strong wind and dust storm area in the middle part of Mongolia. And along with the synoptic systems moved to east, later it influenced Beijing area.

By analyzing the variance of average meteorological elements such as pressure and wind speed (Fig. 3) averaged over six observation stations in Beijing from 3–7 April, it was

found that the average surface pressure of the six stations was lower than 1010 hPa. During the period of 3 April, 16:00 Z to 6 April, 04:00 Z UTC, the average surface wind speed was lower than 2 m/s, which caused pollutions to accumulate and reached heavy pollution levels with an API more than 350 on 5–6 April. Consequently, the observed SO₂ and PM₁₀ concentrations at the IAP site were also very high during this period.

3 Model description and validation

The Models-3/CMAQ, which is developed by the U.S. EPA (Byue, et al., 1999), has already been widely used all over the world especially in North America and East Asia. It has successfully applied to investigate the Asian continental outflow of carbon monoxide, ozone, and aerosol sulfate to the western Pacific Ocean by a number of serious researches (Zhang et al., 2002; Uno et al., 2005; Yamaji et al., 2006).

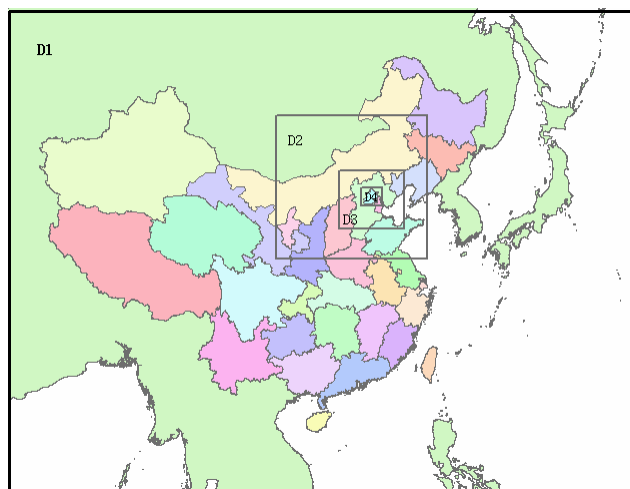
3.1 Model setup

In this study, we use the NCAR/Penn State Mesoscale Model (MM5) (Grell et al., 1994), which is based on MRF scheme for the boundary layer process, Grell scheme for convective motion, cloud-cooling scheme for radiation, and some other physical parameterized schemes, to generate the meteorological fields for CMAQ. The outputted meteorological factors are mainly contained of temperature, wind, humidity, pressure, and other parameters requested by CMAQ. The NCAR/NCEP reanalysis data, four times a day with $1^\circ \times 1^\circ$ resolution, were used for initial and boundary conditions in MM5 model.

Four nested grids at the horizontal resolution of 81 km, 27 km, 9 km and 3 km with the same center located at (35.0° N, 116.0° E) were used for MM5 and CMAQ in this

Table 1. Emissions of major anthropogenic species in this study (Unit: ton/year).

Region	VOC	SO ₂	NO _x	CO	PM ₁₀	PM _{2.5}	NH ₃
Beijing Municipality	285 551	211 306	227 311	1 021 790	106 890	53 367	69 068
Tianjin Municipality	270 000	375 876	178 940	737 041	93 497	37 996	49 981
Hebei Province	855 000	1 353 731	686 000	6 806 000	535 055	264 052	846 462
Shanxi Province	401 000	1 467 410	558 000	3 254 000	173 848	66 899	214 950
Shandong Province	1 088 000	1 575 245	812 000	7 339 000	684 824	379 516	1 093 000

**Fig. 4.** The four nested domains used for simulation.

study (Fig. 4). The coarse domain covers East Asia with 83×65 horizontal grids and the second domain with 61×58 grids includes most provinces of North China. The third domain with 79×70 grids consists of Beijing and the surrounding cities, and the focus of the fourth domain with 73×64 grids on urban Beijing. From the first to the fourth domain, the time step is 240 s, 80 s, 26.67 s and 8.89 s. Thirty levels were unequally distributed in the vertical from the ground to a level of 100 hPa in MM5. And there are 12 vertical layers extending from the surface to approximately 12 km in CMAQ, which are reduced from 30 sigma levels of MM5 using the mass-weighted averaging algorithms for reducing the computational costs. Vertical layers of CMAQ are unevenly distributed and surface layer is approximately 35 m. We used the Grell cumulus parameterization, simple ice physics, and the MRF planetary boundary layer in MM5.

The Sparse Matrix Operator Kernel Emissions (SMOKE, 2005) model was applied to prepare the emission inventory and generate gridded emission data for the CMAQ model. There are two emission inventory data are used in this study for SMOKE processing: the regional data from Streets et al. (2003, and personal communications) with 10 km resolution updated from the TRACE_P emissions inventory, and a detailed local source emissions database of Beijing city with

1 km resolution for point, mobile, area, and others sources based on reports of National Projects on Beijing Air Quality Controls. The selected pollutant emissions used in the modeling for Beijing, Tianjin, and the neighboring provinces (Hebei, Shanxi and Shandong) are summarized in Table 1. Figure 5 presents SMOKE output of emission used in the third domain, clearly showing high anthropogenic emissions of CO, NO_x and SO₂ distributed in urban area of Beijing and Tianjin.

The selected simulation period for analysis is from 2 to 8 April 2005, with 10-day spin-up from 23 March. First three sensitivity runs were conducted to test the uncertainties of NBS emissions impact on Beijing's pollution. In three cases, emissions of local Beijing sources with unchanged, Case 1 with emissions of NBS keeping unchangeable (base run, BR); Case 2 with reduced 50% emissions of NBS, and Case 3 with added 50% emissions of NBS. And then a control run (CR) with only NBS was performed and compared with BR to evaluate the impact of NBS on air pollution in Beijing. All of these simulations were conducted with the same meteorological fields produced by MM5.

3.2 Meteorological comparison of model results with observations

The rationality of MM5's meteorological simulations of this case are exhibited by several statistical parameters such as MB (the Mean Bias), NMB (the Normalized Mean Bias), NME (the Normalized Mean Error), RMSE (the Root Mean Square Error) and R (the Correlation Coefficient) obtained by traditional statistical methods.

The comparison of meteorological modeling and observation finds out that MM5 simulation is accredited in this study. Six meteorological observation stations which are located in Beijing urban area are used to evaluate model results. They are chosen for their representative and consecutive. The average value of the six sites' observation is compared with the meteorological model output such as temperature (T), relative humidity (RH), wind speed (WS), direction (WD) and pressure (P) in Table 2. The correlation coefficients (R) between the observed T, RH, P, WS and WD, and the simulated ones are high and respectively 0.73, 0.71, 0.92, 0.74 and 0.50. So the meteorological MM5 model has the ability to

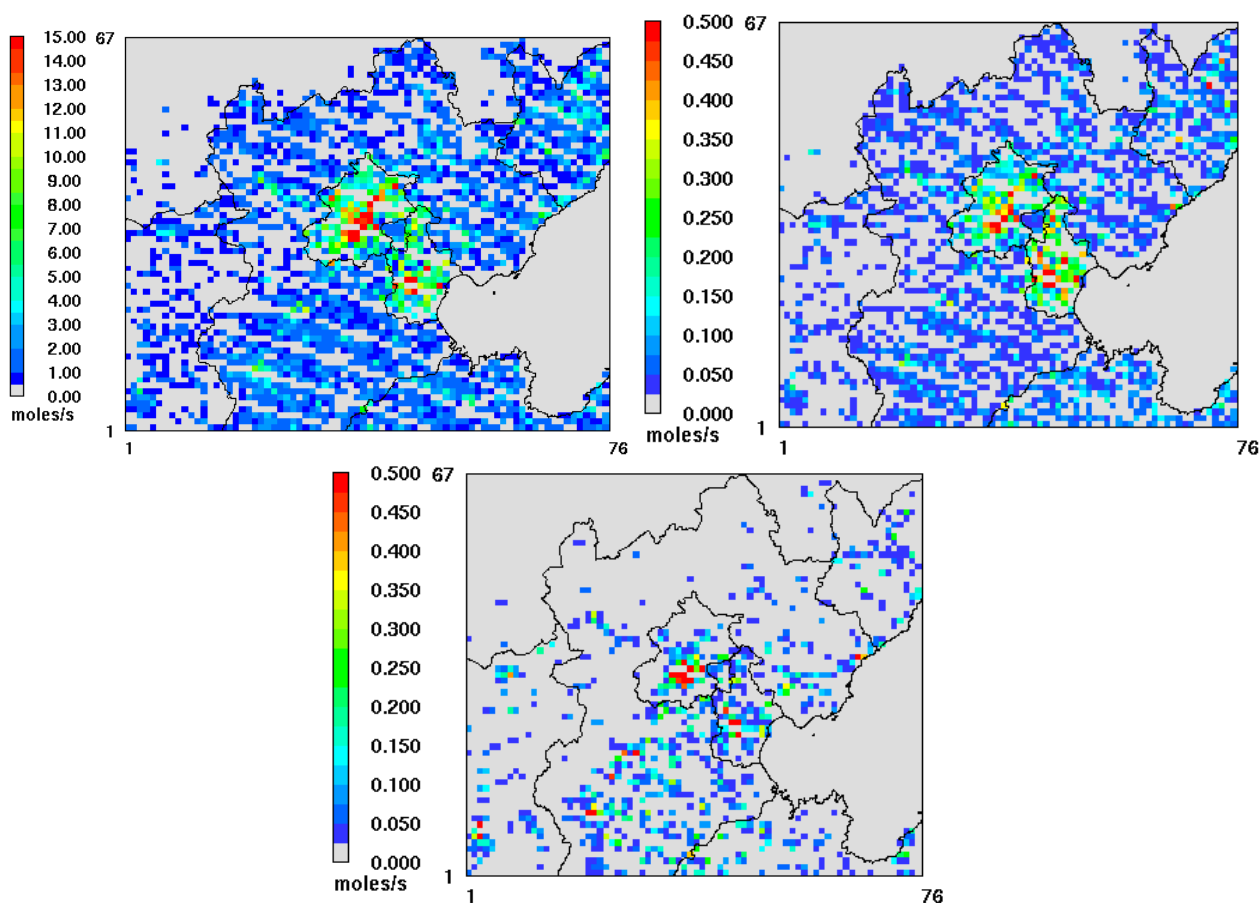


Fig. 5. Spatial distribution of anthropogenic (a) CO, (b) NO_x and (c) SO₂ emission intensity (moles/s) in the third domain with horizontal resolution of 27 km.

Table 2. Statistical summaries of the meteorological comparisons of the model results with observations of average temperature (T), relative humidity (RH), wind speed, direction and pressure, of six observations sites in Beijing.

	Obs.	Sim.	MB	NMB(%)	NME(%)	RMSE	R
T (°C)	15.6	15.5	-0.1	-0.8	21.4	3.8	0.73
RH (%)	46	49	2.2	4.9	26.2	15.4	0.71
Pressure (hPa)	1005	1006	0.4	0.1	0.3	3.7	0.92
Wind speed (m/s)	8.9	13.6	1.4	56.4	58.2	1.8	0.74
Wind direction (°)	170	159	-10.8	-6.4	34.0	73.5	0.50

simulate the different meteorological variables. From other statistical parameters data, such as MB, NMB, NME and RMSE, we can see that the model can better simulate the characteristics of T, RH and P factors. The NMB and NME values for WS are higher than 50%; this indicates that the model slightly overestimates the wind speed. The MB and RMSE are low and the R value is higher for WS, so the model results of WS are reliable.

3.3 Comparison of model results with observation and MODIS AOD

Figure 6 shows the comparison of simulated SO₂ and PM₁₀ concentrations with observed data at 47 m (SO₂) and 8 m (PM₁₀) height of IAP tower. The model simulated the magnitude and the diurnal patterns of SO₂ and PM₁₀ well. The correlation coefficient of SO₂ and PM₁₀ between the model and observation is higher than 0.70 and 0.60, respectively. The

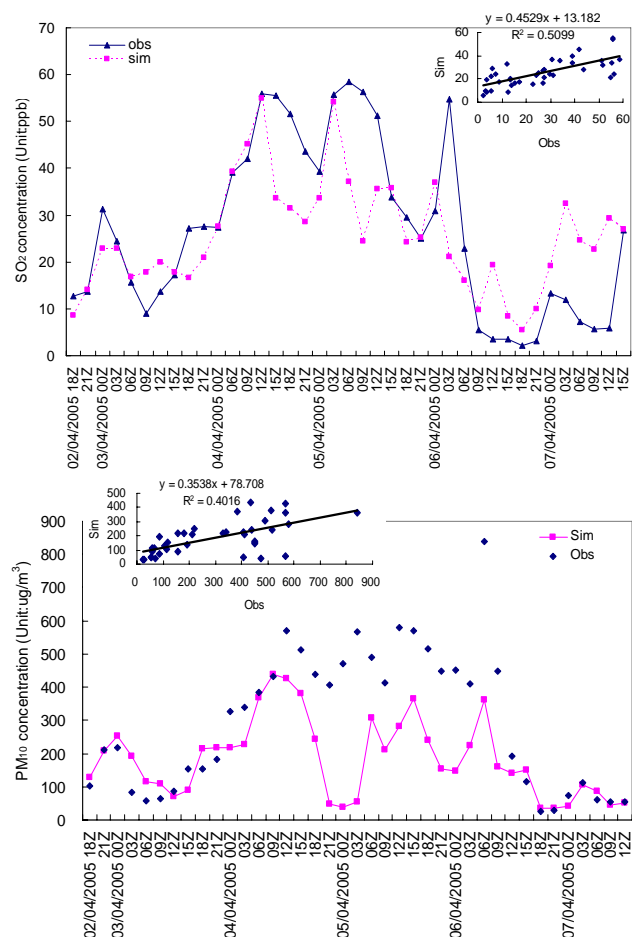


Fig. 6. Comparison of observed and simulated SO₂ (a) and PM₁₀ (b) concentration of three hour average at IAP site from 2 April to 7 April 2005 (UTC).

model can also simulate the variations of the particles during this episode, but the simulation generally under predicted particle concentrations due to uncertainties in emissions and a lack of dust emission module in the current CMAQ system.

The model results are also compared with MODIS optical depth data. Satellite remote sensing provides a means to derive aerosol distribution at global and regional scales. The Moderate Resolution Imaging Spectroradiometer (MODIS) aboard two of NASA's Earth Observing System (EOS) satellites (Terra and Aqua) began a new era of aerosol remote sensing over land (Chu et al., 2003). Aerosol optical depth (AOD) products from MODIS have been verified by the sun-photometer observations in some regions of China (Li et al., 2003) and used to research air pollution in and around Beijing areas (Li et al., 2005). The researched results show that the MODIS AOD data have high accuracy and can better describe the regional and urban pollution distribution. Figure 7a shows the spatial distribution of MODIS Level 2 prod-

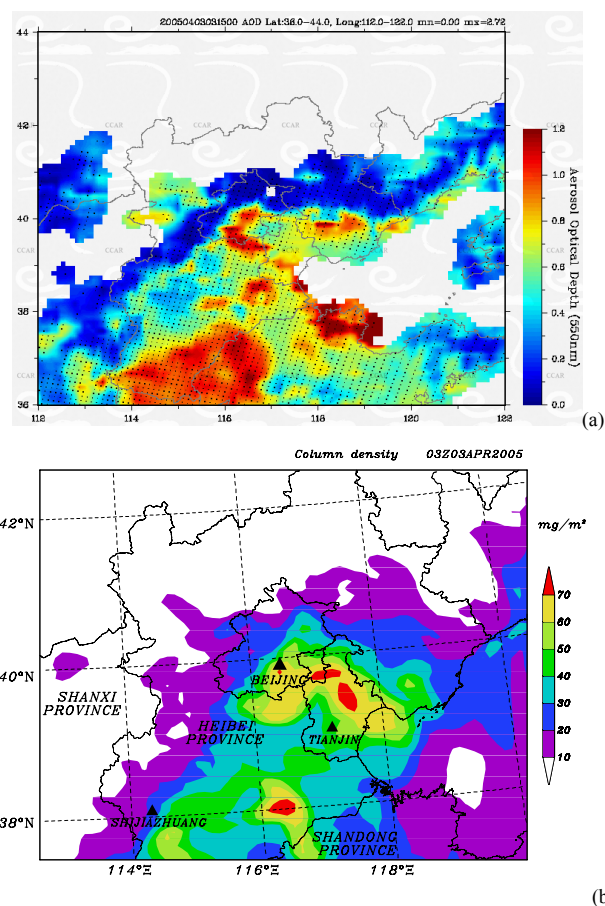


Fig. 7. (a) Aerosol Optical Depth spatial distributions by MODIS at 03:15Z on 3 April 2005 (b) Model simulated PM₁₀ column density of Beijing and surrounding areas at 03:00 (UTC) on 3 April 2005.

ucts of AOD (Version 4) with 10 km nadir spatial resolution in and around Beijing areas on 3 April 2005, 03:15Z. Data over ocean parts are not used for larger uncertainty is usually caused by high concentration of sediment. Regions in the north part of the figure are also blank for the surface is too bright (not covered with dense vegetation) to get precise aerosol retrieval. Figure 7b is the model simulated PM₁₀ column density of Beijing and surrounding areas on 3 April 2005, 03:00Z. Compared Fig. 7a with Fig. 7b, it can be seen that at in most areas the distributions patterns of simulated column density are consistent with MODIS AOD. In the coastal areas of Bo Hai Bay, about 200 km away from Beijing, MODIS observed larger AOD values than our model. The difference between the two sources is mainly due to MODIS AOD retrieval errors near estuary and coastal areas with high sediment concentration in the water (Li et al., 2003). AOD values in the Beijing area corresponding to our model grids are derived from MODIS level 2 products according to the nearest valid data within the scope from 39° N to 40.5° N, and from 115.5° E to 117° E. The

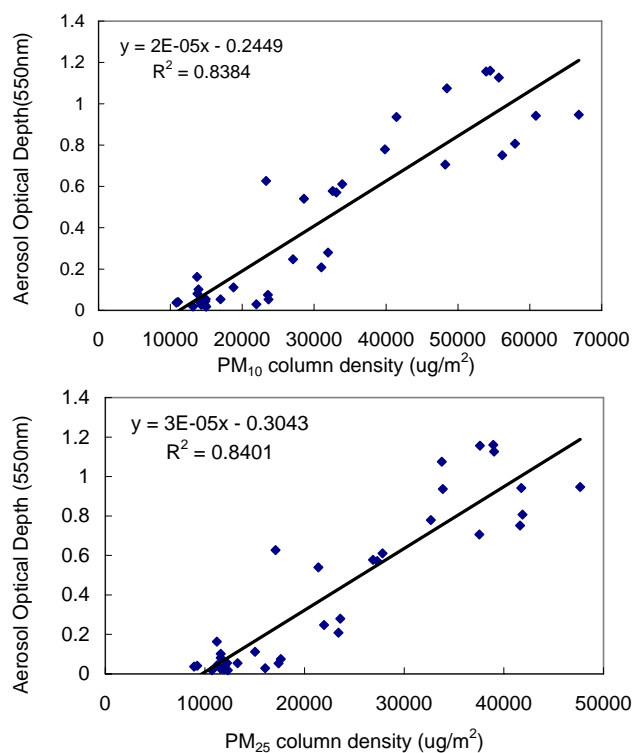


Fig. 8. Correlation coefficient of simulated PM₁₀ (a) and PM_{2.5} (b) column density with Aerosol Optical Depth.

correlation coefficient between MODIS AOD and model results are higher than 0.90 (R-squared \sim 0.84) (Fig. 8).

4 Results and discussions

4.1 Temporal and spatial distribution of simulated PM₁₀ concentration

Figure 9a shows the simulated altitude-time cross sections of PM₁₀ concentrations at the IAP site. From 4–6 April of 2005, PM₁₀ concentrations below 200 m were very high, especially at night, which corresponded to the lower atmospheric boundary layer (ABL) height (Fig. 9b) from 12:00 Z to 00:00 Z (UTC), preventing atmospheric pollutants from dispersing away. Also during this period the surface wind speed was very weak especially from 4 April, 9:00 Z to 5 April, 00:00 Z and from 5 April, 12:00 Z to 6 April, 00:00 Z (UTC) with a wind speed of less than 1.0 m/s (Fig. 3).

4.2 Sensitive models results of NBS emissions

Two sensitivity runs are conducted and compared with the base run to show the correlation of NBS emissions with air pollution in Beijing. Timely variance of PM₁₀ and PM_{2.5} concentration along Beijing under case 1 (base case), case 2,

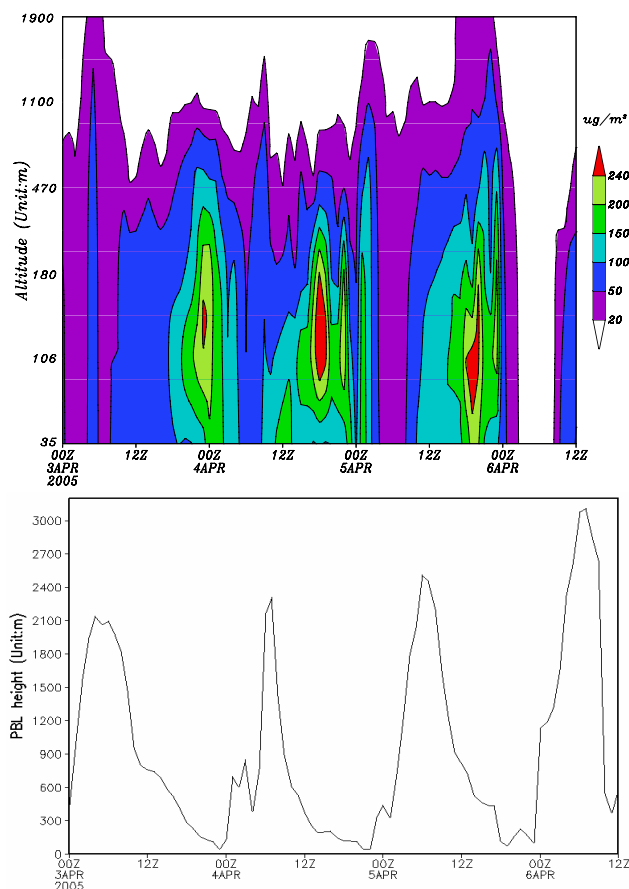


Fig. 9. Vertical distributions of simulated PM₁₀ concentration and calculated atmospheric boundary layer (ABL) height by MM5 along IAP site (a) Altitude-time cross sections of PM₁₀ concentrations (b) ABL height (m).

case 3, and the difference between the cases for PM₁₀ and PM_{2.5} are shown in Fig. 10. Figure 10 shows that the particle concentration variance patterns are similar under the reduced 50% emissions of NBS, added 50% emissions of NBS and base conditions during the episode. The differences between case 1 and case 2, and case 3 and case 1 are nearly the same. It shows that the impact of NBS emissions on air pollution in Beijing is almost linear.

4.3 NBS contributions to SO₂ and particle in Beijing during the episode

Figure 11 represents the simulated SO₂ and PM₁₀ concentrations in the fourth nested domain (3 km) at 02:00 Z, on 5 April. The horizontal distributions of SO₂ and PM₁₀ are shown in Figs. 11a and c with all sources and in Figs. 11b and d for only NBS, respectively. Figure 11a shows that SO₂ concentrations at most urban Beijing are higher than 30 ppb, peaking around at 110 ppb in the west of urban Beijing. Compared with Fig. 11b, the contribution of NBS to

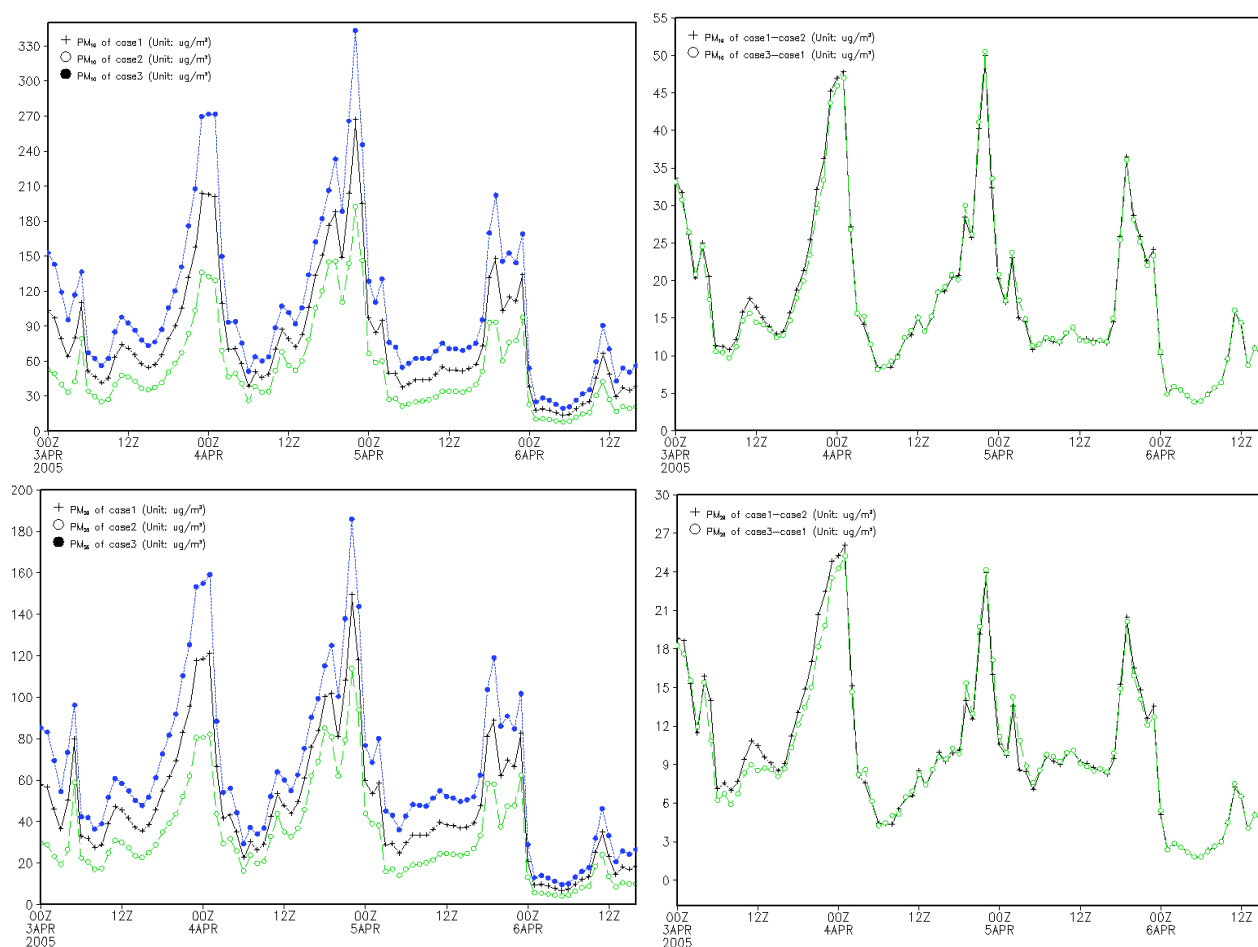


Fig. 10. Comparison of simulated temporal variance of PM₁₀ (a) and PM₂₅ (c) concentration along Beijing under Case 1 (base case), Case 2, Case 3, and the difference between the cases for PM₁₀ (b) and PM₂₅ (d) from 3 April 00Z to 6 April 2005 12Z (UTC). Case 1 with emissions of NBS keeping unchangeable (base run, BR); Case 2 with reduced 50% emissions of NBS, and Case 3 with added 50% emissions of NBS.

SO₂ in most areas of urban Beijing is about 6 ppb, with higher contributions in southeast area. The patterns of PM₁₀ concentrations are similar to those of SO₂, with a maximum of more than 210 $\mu\text{g}/\text{m}^3$ at west of urban Beijing (Fig. 11c). The contribution of NBS for PM₁₀ concentration in Beijing is about 10–15 $\mu\text{g}/\text{m}^3$ (Fig. 11d).

The following formula was used to calculate the percentage contributions (PC) from NBS to PM₁₀, PM_{2.5}, SO₂ and soil (soil particle with diameter between 2.5 μm and 10 μm):

$$\text{PC} = C(\text{CR}) \times 100 / C(\text{BR})$$

where $C(\text{CR})$ and $C(\text{BR})$ are the simulated concentrations of PM₁₀, PM_{2.5}, SO₂ or Soil, with BR run (all sources) and CR run (only NBS), respectively.

Using the results from the two runs, the distribution patterns of PC for PM₁₀ and PM_{2.5} were obtained. PM₁₀ and PM_{2.5} generally show similar patterns, indicating the same source regions for both of them. The contributions of NBS

are variously spatially distributed over urban Beijing. In order to evaluate the effect of NBS on air pollution in different area of Beijing during this episode, the urban area is divided into five regions such as northwest (NW), southwest (SW), center, northeast (NE), southeast (SE). The percentage contribution of NBS to each region is calculated by using weighted mean method and summed up in Table 3. The average percentage (AVE) for all urban area is the average of these five parts. From the table, we can see that the average contribution of NBS to PM_{2.5} is 39%, with maximum up to 53% at the northwest of urban area and only 15% at southwest. The spatial distribution pattern of contribution ratio for PM₁₀ is very close to PM_{2.5}, but with slightly less than PM_{2.5}. The maximum weighted average contribution to PM₁₀ is 43% at the northwest of Beijing urban area and the minimum is 10% at the southwest. And the average contribution for PM₁₀ is 30%. The average contribution of NBS for SO₂ is 18% with the maximum up to 26% at the southeast of

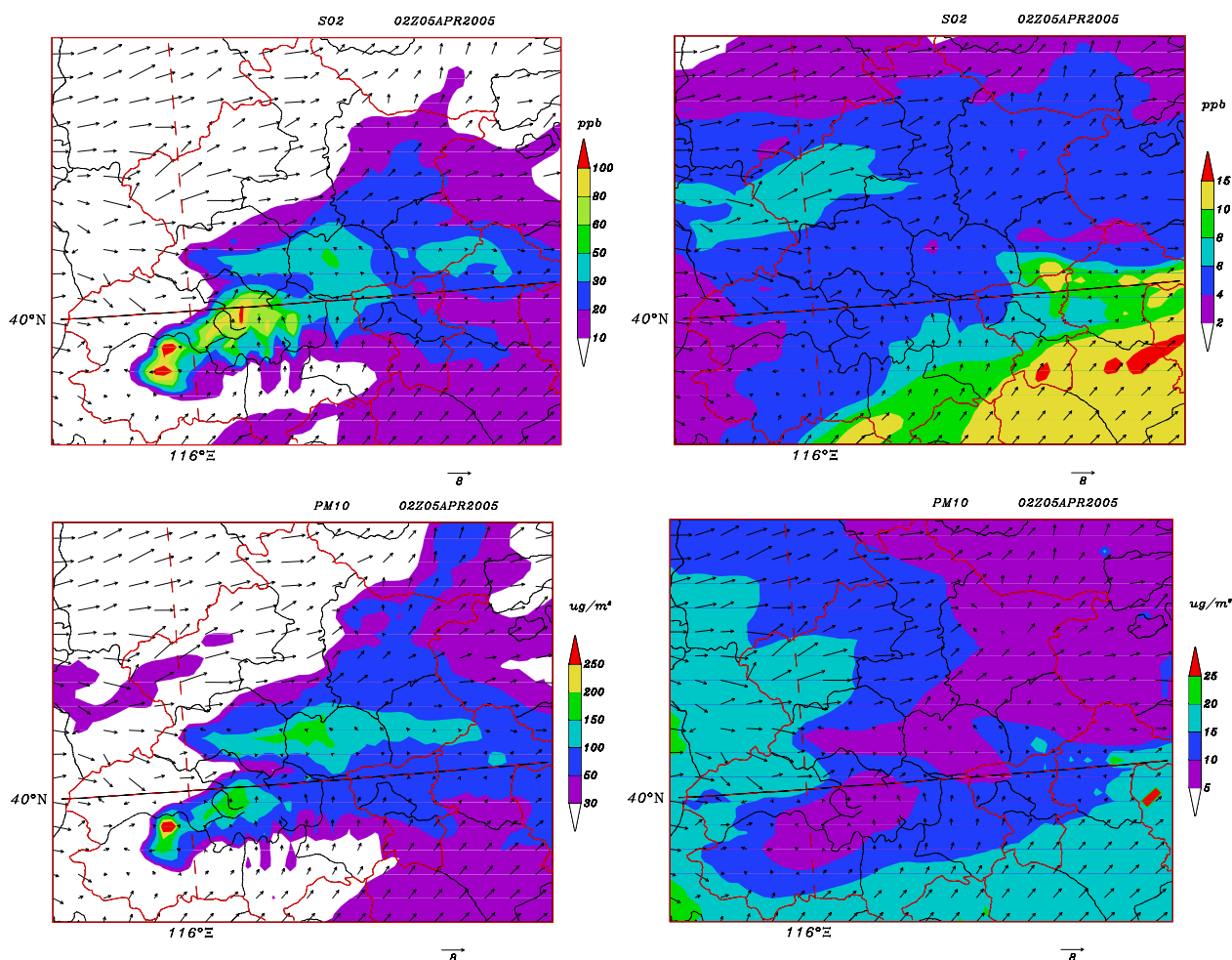


Fig. 11. Horizontal distributions of simulated SO_2 and PM_{10} concentrations over Beijing area at 02:00 (UTC) on 5 April 2005. (a) For SO_2 of all sources; (b) for SO_2 of sources outside Beijing; (c) same as (a), but for PM_{10} ; (d) same as (b), but for PM_{10} .

urban area and the minimum 13% at the northwest. For soil, the spatial distribution pattern of contribution is very like that of SO_2 , and the average contribution of NBS to soil is 10% with the maximum at the southeast and the minimum at the northwest.

Therefore, for $\text{PM}_{2.5}$ and PM_{10} , the order of contribution of NBS from big to small is NW, SE, NE and SW. But for SO_2 and soil, it is different from that of PM_{10} , the order from big to small is SE, SW, NE and NW. Table 3 also indicates that the average contributions of NBS are different for each pollutant, with the order of $\text{PM}_{25} > \text{PM}_{10} > \text{SO}_2 > \text{soil}$ during this episode.

It can also be elucidated from the table that the NBS contribution of fine particles is larger than that of coarse particles to Beijing background PM levels as the fine particles are much easier to be long-range transported.

4.4 Pathways of NBS transporting to Beijing

The vertical altitude-time cross section and the west-east cross section of PM_{10} flux along Beijing are shown in Figs. 12 and 13, respectively. PM_{10} flux is calculated by the following formula:

$$\text{flux} = C \times V$$

Where C is the simulated concentration of PM_{10} and V is the wind velocity.

From 3 to 6 April, strong PM_{10} fluxes were calculated coming from the southwest to the northeast of Beijing, especially during nighttime. In order to clearly reveal the correlation between the NBS contribution to PM_{10} in Beijing, the temporal variance of PM_{10} concentration, PM_{10} flux and the contribution of NBS along Beijing are shown in Fig. 14. From Fig. 14, we can see that the peak PM_{10} flux was seen at about 14:00 Z on 5 April 2005, which corresponded to the

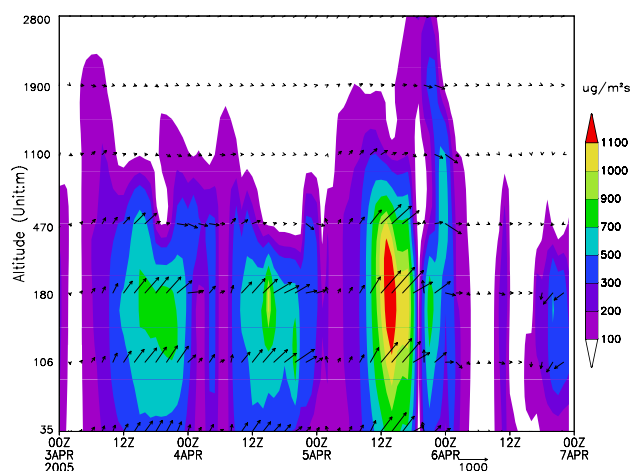


Fig. 12. Altitude-time cross sections of PM_{10} flux (Unit: $\mu\text{g m}^{-2} \text{s}^{-1}$) along Beijing from 3 April to 7 April 2006.

Table 3. The weighted average percentage contributions of NBS to SO_2 and particle in urban Beijing.

Area	NW	SW	Center	NE	SE	AVE
$\text{PM}_{2.5}$	53%	15%	35%	44%	50%	39%
PM_{10}	43%	10%	25%	34%	37%	30%
SO_2	13%	18%	15%	17%	26%	18%
soil	5%	10%	10%	10%	15%	10%

peak value of contribution percentage to PM_{10} at this time. The peak PM_{10} flux was up to $1100 \mu\text{g m}^{-2} \text{s}^{-1}$ at an altitude of about 200 m and correspondingly the peak contribution ratio of NBS to PM_{10} was up to 60%.

Figure 14 shows that the PM_{10} flux peaks at the same time as the maximum of the contribution to PM_{10} occurring, with a delayed PM_{10} concentration peak. At 18:00 Z on 3 April, 15:00 Z on 4 April and at 14:00 Z on 5 April, the contribution percentage of NBS was very large with a strong PM_{10} flux imported from surrounding areas into Beijing, which induced the high PM_{10} concentration at later time (at 23:00 Z on 3 April, 20:00 Z on 4 April and at 20:00 Z on 5 April). It indicates that the contribution of surrounding sources plays a very important role in forming high pollution concentration in Beijing during this episode.

From the horizontal distribution chart of PM_{10} fluxes (Fig. 15), it can be seen that strong particle fluxes were transported from the southwest area of Beijing by southwest wind. At 06:00 Z on 5 April, the flux center located at the boundary of Hebei province from Shanxi province. At 10:00 Z on 5 April, the high flux center moved to the northeast of Shijiazhuang. Later, the flux peak reached Beijing at about 14:00 Z. Therefore, the dominant transport pathway for NBS effecting Beijing was by the west and southwest flow dur-

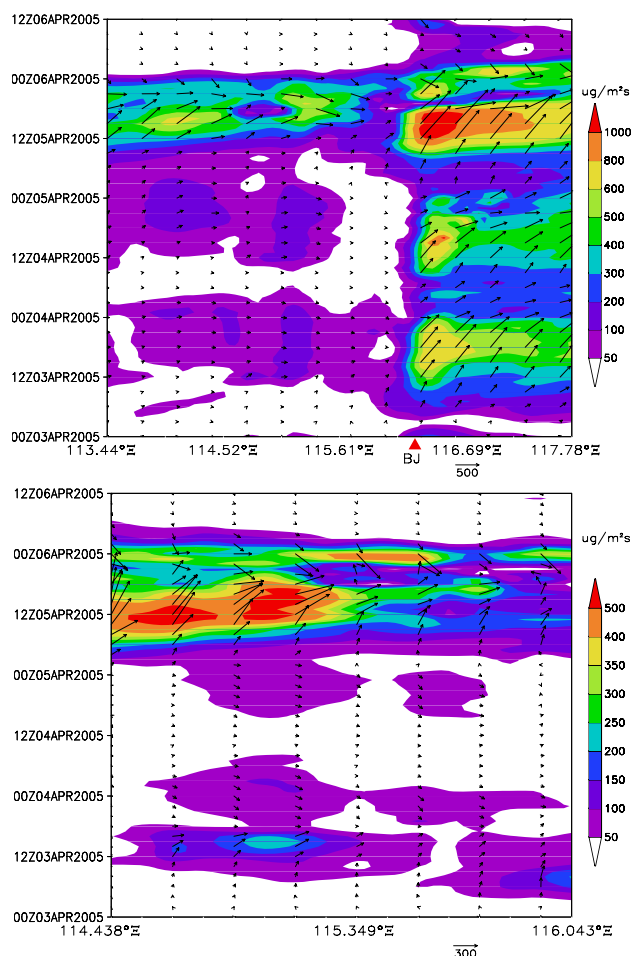


Fig. 13. (a) West-east cross section of PM_{10} flux (Unit: $\mu\text{g m}^{-2} \text{s}^{-1}$) along Beijing from 3 April 00:00 to 6 April 12:00 (UTC), (b) Cross-section of PM_{10} flux (Unit: $\mu\text{g m}^{-2} \text{s}^{-1}$) along SW-NE pathway from 3 April 00:00 to 6 April 12:00 (UTC).

ing this episode. Meteorological analysis (Sect. 2.3) shows that during this period, Beijing was controlled by southwest wind on 850 hPa and surface, and located at low pressure on groove. This type of meteorological pattern offers a convenient condition for sources located at southwest region to be transported into Beijing. And also after 5 April, the floating dust storms brought by a strong Mongolia cyclone across north China strengthened this process.

5 Summary

The Models-3/CMAQ coupled with MM5 and SMOKE module was applied to evaluate the contributions of NBS to high SO_2 and particles pollutants in Beijing during 3–7 April 2005. Comparisons of model results with observed data at the IAP tower and MODIS AOD showed that the model captured the main observed features of this serious pollution

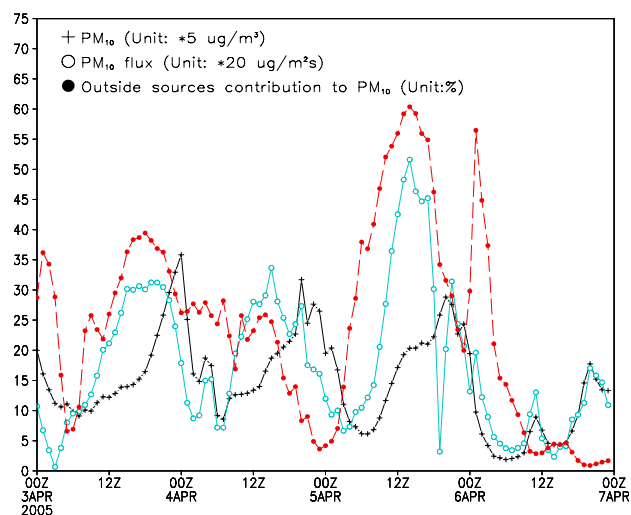


Fig. 14. Temporal variance of PM_{10} concentration, PM_{10} flux and contribution of sources outside Beijing to PM_{10} along Beijing (UTC).

episode. The correlation coefficient for SO_2 between the model and observed data was averaged around 0.70 while the correlation coefficient for the modeled column loading of particles and MODIS AOD reached 0.90. The spatial distributions of SO_2 and PM_{10} simulated by the model showed that high concentrations were transported from southwest area of Beijing to the northeast by west and southwest winds. This was also confirmed with synoptic chart analysis and transport flux analysis during this period. The sensitivity studies indicated that NBS contributions varied temporally and spatially for certain pollution species. For $\text{PM}_{2.5}$, the maximum weighted average contribution was up to 53% at the northwest of urban area, the least was about 15% at southwest. For PM_{10} , spatial distribution and temporal variance of contributions were close to those of $\text{PM}_{2.5}$, but slightly less. Model results also revealed that NBS contributes more to fine particle than coarse particles. The average contributions of SO_2 and soil were 18% and 10%. The calculated transport flux results show that the PM_{10} flux peak and percentage contribution peak of NBS to PM_{10} happened almost simultaneously with a delay the PM_{10} concentrations.

This study presents a well behaved modeling system for investigating the Beijing urban pollutions and source contributions. For a comprehensive analysis of Beijing air pollutions, more observational data are needed for model validations and longer simulations are required to obtain the seasonal variations of NBS contributions. It is noticed that although this method emission switch on/off method is widely used to estimate the contribution of target emissions by most of modellers (Streets et al., 2006), it may cause significant bias/errors to estimate the non-linear source-reception relations. In a nonlinear system with many factors involved, the total impacts of all factors are not linearly additive of the

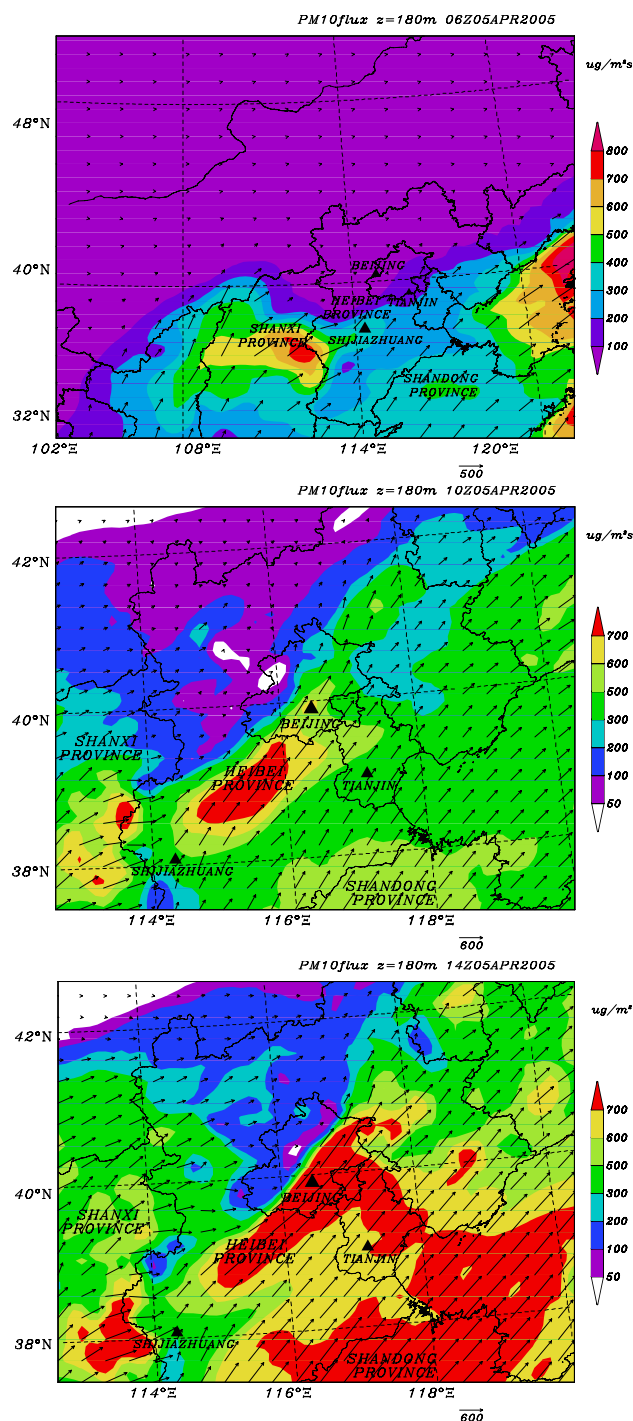


Fig. 15. Horizontal distributions of PM_{10} fluxes at altitude about 180 m at (a) 06:00 Z, (b) 10:00 Z, and (c) 14:00 Z on 5 April (Unit: $\mu\text{g m}^{-2} \text{s}^{-1}$).

contribution from each factor. The contribution of one factor in presence of other factors can be determined by taking the difference between a model run including all factors

and another model run with the factor of interest excluded. For SO₂, it is probably fine, but for PM, there are too many species with complex non-linear reaction and evolution to make uncertainty quite large. The application of the current modeling results for control should be careful but it is an important first step to address the concerns on the regional air quality control needs for Beijing Olympic 2008.

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