



Impact of relative humidity and particles number size distribution on aerosol light extinction in the urban area of Guangzhou

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Abstract. In the urban area of Guangzhou, observations on aerosol light extinction effect were conducted at a monitoring site of the South China Institute of Environmental Sciences (SCIES) during April 2009, July 2009, October 2009 and January 2010. The main goal of these observations is to recognise the impact of relative humidity (RH) and particles number distribution on aerosol light extinction. PM_{2.5} was sampled by Model PQ200 air sampler; ions and OC/EC in PM_{2.5} were identified by the Dionex ion chromatography and the DRI model 2001 carbon analyser, respectively; particles number size distribution was measured by TSI 3321 APS, while total light scattering coefficient was measured by TSI 3563 Nephelometer. Chemical composition of PM_{2.5} was reconstructed by the model ISORROPIA II. As a result, possible major components in PM_{2.5} were (NH₄)₂SO₄, Na₂SO₄, K₂SO₄, NH₄NO₃, HNO₃, water, POM and EC. Regarding ambient RH, mass concentration of PM_{2.5} ranged from 26.1 to 279.1 μg m⁻³ and had an average of 94.8, 44.6, 95.4 and 130.8 μg m⁻³ in April, July, October and January, respectively. With regard to the total mass of PM_{2.5}, inorganic species, water, POM, EC and the Residual accounted for 34–47 %, 19–31 %, 14–20 %, 6–8 % and 8–17 %, respectively. Under the assumption of “internal mixture”, optical properties of PM_{0.5–20} were estimated following the Mie Model. Optical refractive index, hygroscopic growth factor and the dry aerosol density required by the Mie Model were determined with an understanding of chemical composition of PM_{2.5}. With these three parameters and the validated particles number size distribution of PM_{0.5–20}, the temporal vari-

ation trend of optical property of PM_{0.5–20} was estimated with good accuracy. The highest average of $b_{\text{ep,pm0.5–20}}$ was 300 Mm⁻¹ in April while the lowest one was 78.6 Mm⁻¹ in July. Regarding size distribution of $b_{\text{ep,pm0.5–20}}$, peak value was almost located in the diameter range between 0.5 and 1.0 μm. Furthermore, hygroscopic growth of optical properties of PM_{0.5–20} largely depended on RH. As RH increased, $b_{\text{ep,pm0.5–20}}$ grew and favoured a more rapid growth when aerosol had a high content of inorganic water-soluble salts. Averagely, $f_{\text{bep,pm0.5–20}}$ enlarged 1.76 times when RH increased from 20 % to 90 %. With regard to the temporal variation of ambient RH, $f_{\text{bep,pm0.5–20}}$ was 1.29, 1.23, 1.14 and 1.26 on average in April, July, October and January, respectively.

1 Introduction

Aerosol pollution affects radiation budget of the Earth-Atmosphere system by its light extinction. As a result, the global climate changes dramatically (Seinfeld and Pandis, 2006). On the basis of the Mie Model (Bohren and Huffman, 1998), this effect of light extinction can be quantified with the knowledge of single particle light extinction efficiency and particles number size distribution.

Understanding chemical composition is essential to determine the single particle light extinction efficiency. Scientists have developed various kinds of instruments to distinguish the chemical components such as inorganic salts,

organic carbon (OC), element carbon (EC) and crustal elements. Moreover, water-soluble inorganic salts and water-soluble fraction of organic matters absorb water when relative humidity (RH) increases. Subsequently, numerical models like ISORROPIA II (Fountoukis and Nenes, 2007), E-AIM (Wexler and Clegg, 2002) and “Sea Salt” (Eichler et al., 2008) have been developed to recognise the chemical and physical forms of species in aerosol particle at the equilibrium state. However, few details about determining the optical refractive index (ORI) of aerosol particle based on these models were reported.

In Pearl River Delta (PRD) region of China, air visibility degradation in recent years due to an enhancing effect of light extinction made the public and scientists focus attentions on aerosol pollution. And the phenomenon of this pollution was reported in lots of studies (Wang et al., 2003; Bergin et al., 2004; Louie et al., 2005; Wu et al., 2005; Deng et al., 2008a, b; Tie and Cao, 2009). Furthermore, a series of regional integrated field experiments for investigating this pollution were carried out during 2004 and 2008, which mainly concerned:

- chemical transformations among air pollutants and the influence of regional meteorological conditions on pollution episode (Fan et al., 2008; Su et al., 2008; Zhang et al., 2008a, b; Fan et al., 2011);
- aerosol size-resolved chemical composition and the potential pollutant sources (Gnauk et al., 2008; Liu et al., 2008a, b; Jung et al., 2009; Xiao et al., 2011; Yu et al., 2010; Yue et al., 2010);
- mixture state between EC and the other non-light-absorption species in particles (Cheng et al., 2006, 2008b);
- aerosol optical properties, their hygroscopicity and radiative direct forcing (Andreae et al., 2008; Cheng et al., 2008a, b; Eichler et al., 2008; Liu et al., 2010).

Besides the measurement by optical instruments, the Mie Model served as quite an important tool in these experiments because it provided information on the influence of chemical composition and particles number size distribution on aerosol optical properties.

Guangzhou is one of the mega cities in PRD region, while it is also one of the monitoring sites in the field experiment mentioned above. For the purpose of an updated and complementary study on aerosol light extinction in the urban area of this city, four months’ observation on chemical composition and particles number size distribution were carried out at the monitoring site of SCIES (South China Institute of Environmental Sciences) during April 2009, July 2009, October 2009 and January 2010 which represented spring, summer, autumn and winter, respectively. In the light of measurement techniques and numerical models adopted in the previous field experiments, current study tries to reconstruct aerosol chemical composition following the ISORROPIA II

model at first, then to estimate the extent of light extinction effect with a practical method that was based on the result derived from APS (Aerodynamic Particle Sizer) and PM_{2.5} sampling. RH dependence of aerosol optical properties will also be discussed.

2 Experiment

2.1 Monitoring site

The monitoring site of SCIES is located in the urban area of Guangzhou, whose geographical coordinates are 23°07' N and 113°21' E. For monitoring air quality influenced by pollutants’ regional transport and local sources’ emission, instruments were all installed on the roof of the building 53 m above the ground. This site was built with a clear vision of over 300 degrees, around which there is a residential area and a park about 500 m northeast of it. There is no big air pollution source within a circumference of 3 km except mobile emissions. A satellite photo depicting the site’s location and its surroundings is illustrated in Fig. 1. So far, data on aerosol samples, gaseous pollutants and meteorological parameters have been recorded over a long period of time, some of which were once reported in a previous study (Tao et al., 2009).

2.2 Sampling and analysis

2.2.1 Aerosol sampling

PM_{2.5} samples were measured by an air sampler (BGI Corporation, Model PQ200) equipped with a cyclone that separates PM_{2.5} particles from the aerosol population and with a vacuum pump that draws air at a flow rate of 16.7 L min⁻¹. The drawn airstream was connected to a 47 mm quartz filter (Whatman, QM-A). Before sampling, the quartz filters were baked at 800 °C for more than 3 h to remove adsorbed organic vapours, and then equilibrated in desiccators for 24 h. Prior to the measurement in ambient, the flow rate of PM_{2.5} sampler was calibrated. Blank filters were collected and used to subtract the positive artifact caused by gas absorption. Totally, 123 daily quartz-filter samples with some blank ones were collected for every 23.5 h (starting at 10:00 LST each day and ending at 09:30 LST the following day) in the four months. The analysis-ready samples were stored in a freezer at about -20 °C in case of particle volatilisation.

In OC/EC analysis, a punch of 0.5 cm² from the collected quartz filter was analysed for eight carbon fractions following the IMPROVE_A thermal/optical reflectance (TOR) protocol by a DRI model 2001 carbon analyser (Atmoslytic Inc., Calabasas, CA) (Cao et al., 2007; Chow et al., 2007). This process produced four OC fractions (OC1, OC2, OC3 and OC4) at 140 °C, 280 °C, 480 °C and 580 °C, respectively, in a helium [He] atmosphere; OP (a pyrolysed carbon fraction) was determined when transmitted laser light attained



Fig. 1. Satellite photo of the monitoring site and surroundings (from Google Earth).

its original intensity after oxygen [O_2] added to that analysis atmosphere; and three EC fractions (EC1, EC2 and EC3) were determined at 580 °C, 740 °C and 840 °C, respectively, in a (2 %) O_2 /(98 %) [He] atmosphere. IMPROVE_TOR OC is practically defined as $OC1 + OC2 + OC3 + OC4 + OP$, while EC is defined as $EC1 + EC2 + EC3 - OP$ (Chow et al., 2007). Inter-laboratory sample comparisons between applying the IMPROVE_TOR protocol and the TMO (thermal manganese dioxide oxidation) approach have shown the differences being lower than 5 % for TC and 10 % for OC and EC (Chow et al., 2007). Average field blanks were 1.8 and $0.1 \mu\text{g m}^{-3}$ for OC and EC, respectively.

In analysis of water-soluble ions, one quarter of the collected quartz filter sample was used to determine the ions' mass concentrations. Four anions (SO_4^{2-} , NO_3^- , Cl^- and F^-) and five cations (Na^+ , NH_4^+ , K^+ , Mg^{2+} and Ca^{2+}) in aqueous extracts from the filter were determined by ion chromatography (Dionex Corp, Sunnyvale, CA, Model Dionex 600). For these extractions, each sample was put into a separate 20 mL vial containing 10 mL distilled-deionised water (18 M Ω resistivity), and shaken first by an ultrasonic instrument for 60 min, then by a mechanical shaker for 1 h for a complete extraction. The extracts were stored at 4 °C in a pre-cleaned tube before further analysis. Cation (Na^+ , NH_4^+ , K^+ , Mg^{2+} and Ca^{2+}) concentrations were determined with a CS12A column (Dionex Corp, Sunnyvale, CA.) and 20 mmol L $^{-1}$ MSA eluent. Anions (SO_4^{2-} , NO_3^- , Cl^- and F^-) were separated by an AS11-HC column (Dionex Corp, Sunnyvale, CA) and 20 mmol L $^{-1}$ KOH eluent. The limits of detection were less than 0.05 mg L^{-1} for both cations and anions. Standard reference materials produced by the National Research Centre for Certified Reference Materials in China were analysed for the purposes of quality assurance. Blank values were subtracted from sample concentrations (Shen et al., 2008).

It is said that there may be artifacts when using quartz filter for $PM_{2.5}$ sampling. However, the high loading of $PM_{2.5}$ together with the damp climate in Guangzhou can block the

Teflon filter easily, which affects the flow rate of the sampler and increases the sampling error. For this reason, there were studies (Shen et al., 2009; Wang et al., 2011) on aerosol sampling in China using the quartz filter. Moreover, though the loss of quartz filter debris may lead to the underestimation of aerosol mass, careful operations in the process of sampling and mass weighing minimised this loss as much as possible in the current study.

The field blanks were determined and the average values of 12 blank filters of Na^+ , NH_4^+ , K^+ , Mg^{2+} , Ca^{2+} , F^- , Cl^- , NO_3^- , and SO_4^{2-} were 0.671 ± 0.091 , 0.002 ± 0.002 , 0.005 ± 0.006 , 0.006 ± 0.007 , 0.052 ± 0.064 , 0.168 ± 0.036 , 0.425 ± 0.094 , 0.077 ± 0.096 and $0.362 \pm 0.082 \text{ mg L}^{-1}$, respectively. Although the blank value of Na^+ , F^- , Cl^- and Ca^{2+} were slightly higher than other species, blank filter was collected every 10 samples and the blank values were quite stable. All results in the present study were blank subtracted. Moreover, the values of ambient samples were significantly higher than the blank value, which can reduce the errors.

2.2.2 Measurement of particles number size distribution

Particles number size distribution of $PM_{0.5-20}$ was measured by APS (TSI Aerodynamic Sizer, Model 3321) with 52 size bins in the diameter range from 0.5 to 20 μm by determining the time-of-flight of an individual particle in an accelerating flow field. To capture dry particles, a drying tube was added in the process of drawing air. Flow rate of 5 L min^{-1} and 5 min data average were set in APS operation.

2.2.3 RH measurement

Ambient RH had been recorded every 30 min by an automatic weather station (VASALA Model QMH102).

2.2.4 Aerosol optical properties measurement

Total light scattering coefficient of aerosol was measured by an integrating Nephelometer (TSI Performance Measurement Tools, Model 3563) in wavelengths of 450 nm, 550 nm and 700 nm, respectively. Nephelometer calibration was performed by carbon dioxide (CO_2) as high-span gas and filtered air as low-span gas. Nephelometer drew ambient air through a temperature-controlled inlet at a flow rate of 20 L min^{-1} . The inner heater controlled the RH of air intake at a level lower than nearly 70 %. The output data were set to be 1 min average, and zero level data was measured continuously for 5 min after each hourly (60 min) sampling.

The aerosol properties and meteorological parameters measured during the four months' observations are summarised in Table 1.

The result of total light scattering coefficient was corrected for Angular Nonidealities following the method stated in a previous study (Anderson and Ogren, 1998) where a linear function of Angstrom exponent $\hat{a}_{450/700}$ was used as a

Table 1. Measured aerosol properties and meteorological parameters.

	Data Set	Relative humidity	Time resolution	Available data in days			
				Apr-09	Jul-09	Oct-09	Jan-10
1	PM _{2.5} mass concentration	40 %	23.5 h	30	31	31	31
2	Total light scattering coefficient (450 nm, 550 nm, 700 nm)	17–73 %*	1min	29	31	31	31
3	Particles number concentration in the range of 0.5–20 μm	dry	5min	30	30	29	31
4	Relative humidity	29–88 %	30min	30	31	31	30

* Stands for the RH detected by a built-in RH sensor in Nephelometer.

correction factor. Furthermore, the results of particles number concentration, total light scattering coefficient and RH were calculated into daily averages to be compatible with daily PM_{2.5} samples. Blank records in these parameters were estimated by a linear interpolation based on those validated ones.

3 Methodology

3.1 Chemical composition reconstruction

On the basis of identified cations (Na⁺, NH₄⁺, Ca²⁺, K⁺, Mg²⁺) and anions (SO₄²⁻, NO₃⁻, Cl⁻), the model ISORROPIA II was introduced to determine the chemical and physical forms of inorganic species and the content of water uptake. Currently, ISORROPIA II was set to solve a “Forward” problem, the result of which was in “Metastable” state as aerosol particle was assumed to be composed of an aqueous supersaturated solution. Accordingly, several aqueous species (Na⁺, HSO₄⁻, SO₄²⁻, etc.) were determined, but to recognise their compound forms is still difficult.

The present study intends to associate these aqueous species into possible compounds according to the five aerosol composition regimes defined in ISORROPIA II. These regimes are “Sulfate Rich (free acid)”; “Sulfate Rich”; “Sulfate Poor, Crustal & Sodium Poor”; “Sulfate Poor, Crustal & Sodium Rich, Crustal Poor” and “Sulfate Poor, Crustal & Sodium Rich, Crustal Rich”. Making use of the cations and anions identified in every PM_{2.5} sample, characteristic parameters R_1 (Ratio of sum of Na⁺, NH₄⁺, Ca²⁺, K⁺, Mg²⁺ to SO₄²⁻), R_2 (Ratio of sum of Na⁺, Ca²⁺, K⁺, Mg²⁺ to SO₄²⁻) and R_3 (Ratio of sum of Ca²⁺, K⁺, Mg²⁺ to SO₄²⁻) defined by ISORROPIA II were calculated. With regard to their values, chemical composition of each PM_{2.5} sample was related to one of the five regimes. Furthermore, one is able to associate those aqueous species into compounds following several principles adopted by the model including:

- there is an electric charge balance in “Metastable” state;
- Na⁺ and K⁺ preferentially associate with SO₄²⁻ before combining with NO₃⁻ and Cl⁻ because SO₄²⁻ is less volatile than NO₃⁻ and Cl⁻;

- SO₄²⁻ associates with Na⁺. If any remain, then it associates with NH₄⁺. The same assumption applies for HSO₄⁻;
- gas species NH₃, HCl and HNO₃ dissolve through the equilibria of the supersaturated aerosol system.

Consequently, possible inorganic species were determined in the form of NH₄HSO₄, (NH₄)₂SO₄, NaHSO₄, Na₂SO₄, NH₄NO₃, NaNO₃, NH₄Cl, NaCl, K₂SO₄, MgSO₄, CaSO₄, KHSO₄, H₂SO₄, NH₃, HCl and HNO₃. It should be noted that the volatilisation effect of NH₄NO₃ and other semi-volatile species was not investigated in the current study.

As the hygroscopicity of water-soluble inorganic salts is considered in ISORROPIA II, the content of water uptake in each PM_{2.5} sample was also determined.

Particulate organic matter (POM) is an important chemical component, which was estimated by the content of OC being multiplied by a factor of 1.6 (Cao et al., 2007). On the other hand, the hygroscopicity of POM is not considered currently because:

- as a whole, the hygroscopic growth of Secondary Organic Aerosol (SOA) was found to be around 1.2 at 90 % RH (Gysel et al., 2007; Stock et al., 2011);
- the hygroscopicity of some extracts from Water Soluble Organic Carbon (WSOC) has been recognised (Gysel et al., 2004). However, there is no WSOC speciation in the present study;
- water uptake by the aged organic aerosol accounted for only a few percent of total water uptake (Bougiatioti et al., 2009; Engelhart et al., 2011);
- unlike the water-soluble inorganic salts, a more accurate RH dependence curve of POM has not been well established;
- organic species have not been included in the ISORROPIA II model.

Since inorganic species, water, POM and EC were recognised, remaining unidentified species were categorised to the “Residual”. EC and the Residual were assumed to have no

Table 2. Summary of the parameters for calculating the EORI and EGF.

Specie	ρ_i (g cm ⁻³)	n_i	k_i	Specie	ρ_i (g cm ⁻³)	n_i	k_i
NH ₄ HSO ₄	1.780	1.473	0.000	Ca(NO ₃) ₂	2.504	1.530	0.000
(NH ₄) ₂ SO ₄	1.760	1.530	0.000	CaCl ₂	2.150	1.520	0.000
NaHSO ₄	2.476	1.460	0.000	Mg(NO ₃) ₂	2.020	1.510	0.000
Na ₂ SO ₄	2.680	1.480	0.000	MgCl ₂	2.325	1.540	0.000
NH ₄ NO ₃	1.725	1.554	0.000	KNO ₃	2.110	1.504	0.000
NaNO ₃	2.261	1.587	0.000	KCl	1.980	1.490	0.000
NH ₄ Cl	1.527	1.639	0.000	NH ₃	0.597	1.000	0.000
NaCl	2.160	1.544	0.000	HCl	1.200	1.250	0.000
K ₂ SO ₄	2.660	1.490	0.000	HNO ₃	1.500	1.400	0.000
MgSO ₄	2.660	1.560	0.000	Water	1.000	1.333	0.000
CaSO ₄	2.610	1.570	0.000	POM	1.400	1.550	0.001
H ₂ SO ₄	1.840	1.430	0.000	EC	1.500	1.800	0.540
KHSO ₄	2.245	1.480	0.000	Residual	2.000	1.580	0.005

hygroscopic growth. In this regard, the content of the Residual was the difference between PM_{2.5} and the sum of the identified species at 40 % RH. Furthermore, mass concentration of PM_{2.5} at any other RH condition can be calculated based on this determined content of the Residual.

3.2 Aerosol optical property estimation

In consideration of the APS measurement and PM_{2.5} sampling in the present study, an assumption of “internal mixture” was introduced into the Mie Model, which considers every chemical component in a particle as homogeneously mixing with each other (Jacobson, 2001; Bond and Bergstrom, 2006; Cheng et al., 2008c).

The EORI represents the “average” ORI of an “internal mixture” particle, which can be calculated with the ORI of each component following mixing rule of Volume-Average (Lesins et al., 2002). The formulas for the EORI are written as Eqs. (1) and (2).

$$n_{\text{eff}} = \left(\sum_i n_i \cdot m_i / \rho_i \right) / \left(\sum_i m_i / \rho_i \right) \quad (1)$$

$$k_{\text{eff}} = \left(\sum_i k_i \cdot m_i / \rho_i \right) / \left(\sum_i m_i / \rho_i \right) \quad (2)$$

In Eqs. (1) and (2), m_i stands for mass concentration of the i -th component in particles, while ρ_i is the density. Respectively, n_i is the real part of ORI of the i -th component, k_i is the imaginary part. Regarding the EORI, n_{eff} is the real part, and k_{eff} is the imaginary part.

Furthermore, the EGF is the “average” hygroscopic growth factor, which can be calculated with Eq. (3) (Eichler et al., 2008).

$$f_{\text{eff}} = \left(\sum_i \varepsilon_i \cdot f_{g,i}^3 \right)^{1/3} = \left(\frac{V_{\text{water}} + V_{\text{dry}}}{V_{\text{dry}}} \right)^{1/3} = \left(\frac{m_{\text{wet}} \cdot \rho_{\text{dry}}}{m_{\text{dry}} \cdot \rho_{\text{wet}}} \right)^{1/3} \quad (3)$$

$$\rho = \left(\sum_i m_i \right) / \left(\sum_i m_i / \rho_i \right) \quad (4)$$

In Eq. (3), f_{eff} is the EGF. ε_i is the volume fraction of the i -th component in aerosol, while $f_{g,i}$ is the hygroscopic growth factor. V_{water} and V_{dry} is the volume of water uptake and dry particle, respectively. The dry aerosol density, ρ_{dry} , in Eq. (3) can be calculated with Eq. (4) where ρ_i is the density of the i -th chemical component (excluding water) and m_i is mass concentration.

Parameters for calculating the EORI and the EGF are summarised in Table 2, which were learned from previous studies (Tang, 1996; Chazette and Lousse, 2001; Sloane, 1986; Haynes, 2011; Seinfeld and Pandis, 2006; Eichler et al., 2008). n_i and k_i in the table are referenced to light wavelength of 550 nm.

According to the Mie Model, b_{sp} (light scattering coefficient) and b_{ep} (light extinction coefficient) can be quantified with Eqs. (5) and (6) (Bohren and Huffman, 1998; Seinfeld and Pandis, 2006), respectively. b_{ap} (light absorption coefficient) is the difference between b_{ep} and b_{sp} . Optical properties including b_{ep} , b_{sp} and b_{ap} to be discussed later are all referenced to light wavelength of 550 nm.

$$b_{\text{sp}} = \sum_j b_{\text{sp},j} = \sum_j \frac{\pi D_j^2}{4} \cdot Q_{\text{sp},j}(D_j, \lambda, \text{EORI}_j) \cdot N_j \quad (5)$$

$$b_{\text{ep}} = \sum_j b_{\text{ep},j} = \sum_j \frac{\pi D_j^2}{4} \cdot Q_{\text{ep},j}(D_j, \lambda, \text{EORI}_j) \cdot N_j \quad (6)$$

In Eqs. (5) and (6), D_j stands for the midpoint Stokes Diameter in the j -th particle size range, while N_j is the number concentration of particles with diameter D_j . $Q_{\text{sp},j}$ represents light scattering efficiency of a single particle with diameter D_j , while $Q_{\text{ep},j}$ represents light absorption efficiency. Theoretically, $Q_{\text{sp},j}$ and $Q_{\text{ep},j}$ are both the function of D_j and the EORI _{j} (the EORI of the particle with diameter D_j) at a given

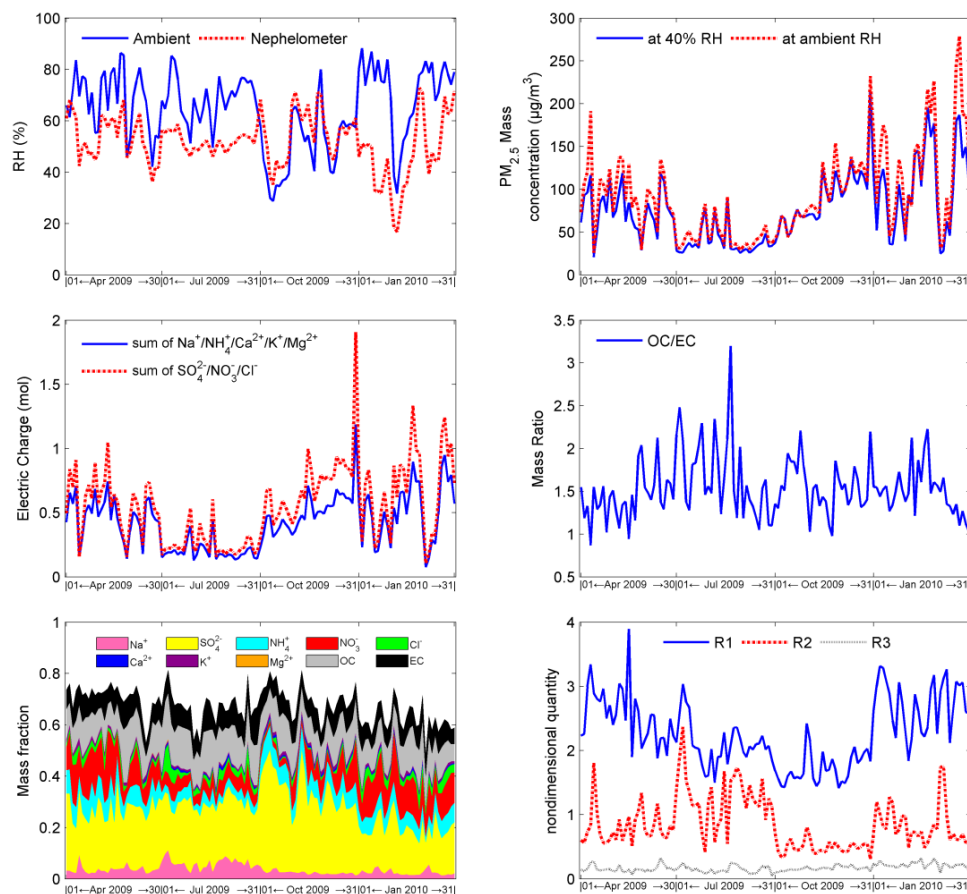


Fig. 2. Temporal variations of RH, PM_{2.5} mass, molarity of electric charge, ratio of OC to EC, components' mass fraction and characteristic values (R_1 , R_2 , R_3).

light wavelength λ (say 550 nm), for which the complicated calculations were referenced to a previous publication (Seinfeld and Pandis, 2006). Regarding the limitation of measurement techniques, the $EORI_j$ was assumed to be equal to the $EORI_{pm2.5}$ which was determined based on chemical composition of PM_{2.5}.

The D_j required by the Mie Model was converted from $D_{a,j}$ of APS size bin with Eq. (7), while the corresponding N_j was derived from APS. Moreover, $f_{eff,j}$ in Eq. (7) represented the hygroscopicity of D_j under an assumption that no change in particles number through the process of hygroscopic growth (Eichler et al., 2008). Sharing the similarity of $EORI_j$, $f_{eff,j}$ was assumed to be equal to the $f_{eff,pm2.5}$.

$$D_j = \frac{D_{a,j}}{\sqrt{\rho_{dry}}} \cdot f_{eff,j}. \quad (7)$$

4 Result and discussion

4.1 Chemical composition of PM_{2.5}

As illustrated in Fig. 2, mass concentration of PM_{2.5} at 40 % RH ranged from 21.0 to 213.6 $\mu\text{g m}^{-3}$ during the four months, and the average of 76.0, 38.6, 89.3 and 103.3 $\mu\text{g m}^{-3}$ were recorded in April, July, October and January, respectively. Figure 2 also shows a coherent temporal variation between the sum of the anions and that of the cations, with the former being sufficient to neutralise the latter. Regarding the content of components in PM_{2.5}, inorganic ions accounted for 33–57 % of total mass; OC was 12–14 %, while EC was about 8 %. The mass ratio of OC to EC ranged from 0.9 to 3.2 and had an average of 1.5 during the four months.

Characteristic values R_1 , R_2 and R_3 were calculated and their temporal variations are illustrated in Fig. 2 as well. PM_{2.5} samples were then related to the five aerosol chemical composition regimes. It should be noted that the electric charge balance in the aerosol system is examined by ISOR-ROPIA II itself, and no unbalance was detected in the current study.

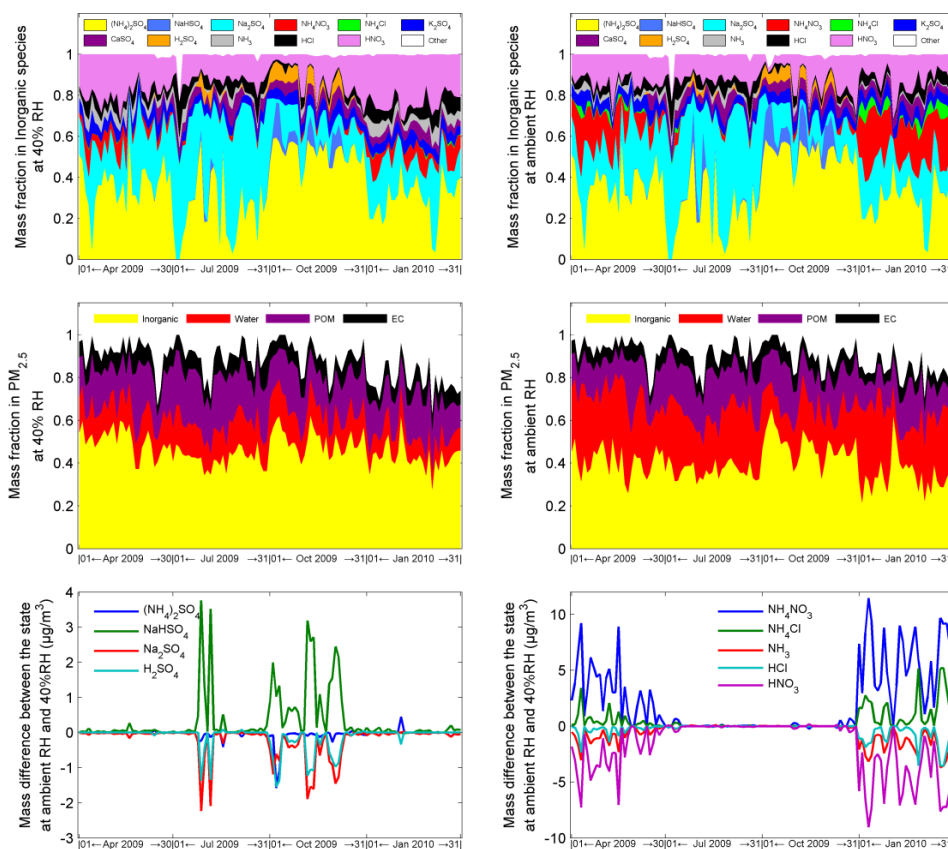


Fig. 3. Temporal variations of chemical composition of $\text{PM}_{2.5}$ at 40 % RH and ambient RH.

Table 3 notes that the prevailing regime was “Sulfate Poor, Crustal & Sodium Poor” in April and January, while “Sulfate rich” in October. These two regimes accounted for about 32 % and 65 % of total samples in July, respectively. During October, a drier climate with strong solar radiation in South China probably accelerated the oxidation process of SO_2 to become SO_4^{2-} , hence the content of sulfate in ambient atmosphere was getting richer.

According to the methodology mentioned earlier, chemical composition of $\text{PM}_{2.5}$ was reconstructed, whose temporal variations are illustrated in Fig. 3. It should be noted that an empirical OC/POM conversion probably leads to an over-estimation of POM besides the errors in measurements and water content calculation. As a result, mass of the Residual in 10 of the 123 samples were calculated to small negatives. To avoid this matter affecting further calculation in the Mie Model, these negative values were assigned to zero. As illustrated in Fig. 3, of the total mass, inorganic species, water, POM, EC and the Residual accounted for 42–51 %, 10–15 %, 17–23 %, 8–9 % and 10–22 %, respectively. Moreover, $(\text{NH}_4)_2\text{SO}_4$, Na_2SO_4 , K_2SO_4 and HNO_3 are the major inorganic species during the four months. In October, the content of NH_4HSO_4 and H_2SO_4 rose for a sulfate rich atmosphere.

NH_4NO_3 usually appeared in April and January when there was not enough HSO_4^- or SO_4^{2-} to neutralise the NH_4^+ .

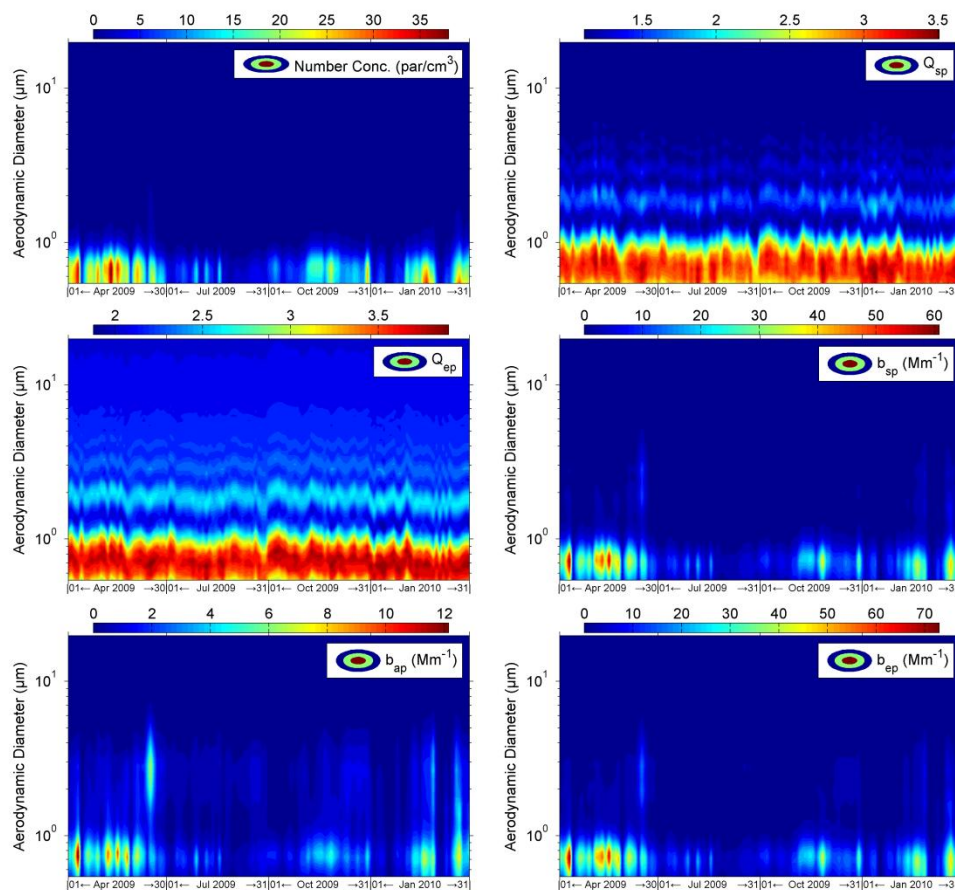
As mentioned above, $\text{PM}_{2.5}$ mass at ambient RH was determined according to its mass at 40 % RH. With regard to the content at 40 % RH, water uptake varied significantly with ambient RH. Compound forms of inorganic species changed as well. As illustrated in Fig. 3, NaHSO_4 , Na_2SO_4 and H_2SO_4 often had significant variations during July and October when sulfate was rich. In April and January, more NH_4^+ , Cl^- and NO_3^- dissociated from dissolved NH_3 , HCl and HNO_3 , hence, NH_4NO_3 and NH_4Cl increased their amounts. At ambient RH condition, mass concentration of $\text{PM}_{2.5}$ ranged from 26.1 to 279.1 $\mu\text{g m}^{-3}$ and had an average of 94.8, 44.6, 95.4 and 130.8 $\mu\text{g m}^{-3}$ in April, July, October and January, respectively. With regard to the composition, inorganic species, water, POM, EC and the Residual accounted for 34–47 %, 19–31 %, 14–20 %, 6–8 % and 8–17 % of total mass, respectively.

4.2 Optical properties of $\text{PM}_{0.5-2.0}$

With the understanding of the aerosol chemical composition reconstructed above, the EORI, EGF and ρ_{dry} required by the Mie Model were determined. Owing to these

Table 3. Quantities of PM_{2.5} samples which were categorised to the five chemical composition regimes.

Condition	Regime	Apr-09	Jul-09	Oct-09	Jan-10
$R_1 < 1$	Sulfate Rich (free acid)	0	0	0	0
$1 \leq R_1 < 2$	Sulfate Rich	3	10	27	1
$R_1 \geq 2, R_2 < 2$	Sulfate Poor, Crustal & Sodium Poor	27	20	4	30
$R_1 \geq 2, R_2 \geq 2, R_3 \leq 2$	Sulfate Poor, Crustal & Sodium Rich, Crustal Poor	0	1	0	0
$R_1 \geq 2, R_2 \geq 2, R_3 > 2$	Sulfate Poor, Crustal & Sodium Rich, Crustal Rich	0	0	0	0

**Fig. 4.** Temporal variations of particles number and optical properties.

three key parameters and particles number size distribution of PM_{0.5–20} from APS, b_{ep} , b_{sp} and b_{ap} of PM_{0.5–20} were calculated following the methodology mentioned earlier. At ambient RH condition, the $EORI_{pm2.5}$ had an average of $1.462-0.037i$, while $EGF_{pm2.5}$ and ρ_{dry} was 1.487 and 1.848 g cm^{-3} , respectively. Subsequently, size distributions of optical properties of PM_{0.5–20} were plotted in Fig. 4 where peak values of particles number and optical properties were almost located in the diameter range between 0.5 and $1.0 \mu\text{m}$. Since the volatilisation loss of semi-volatile species was not investigated in present study, the calculated b_{ep} of PM_{0.5–20} was probably underestimated when the volatilisation effect was important. Moreover, as organic species have

not been considered by the ISORROPIA II, the content of water uptake was slightly underestimated and, therefore, b_{ep} of PM_{0.5–20} was probably underestimated.

4.2.1 Impact of particles number size distribution

In Fig. 5, temporal variation comparison between $b_{sp,pm0.5-20}$ and $b_{sp,total}$ ($b_{sp,total}$ was measured by Nephelometer) was made in consideration of the former one being subset of the latter. It should be noted that the $EORI_{pm2.5}$, $EGF_{pm2.5}$ and ρ_{dry} for $b_{sp,pm0.5-20}$ in this comparison were calculated at the RH condition inside Nephelometer and had the average of $1.473-0.042i$, 1.406 and 1.806 g cm^{-3} ,

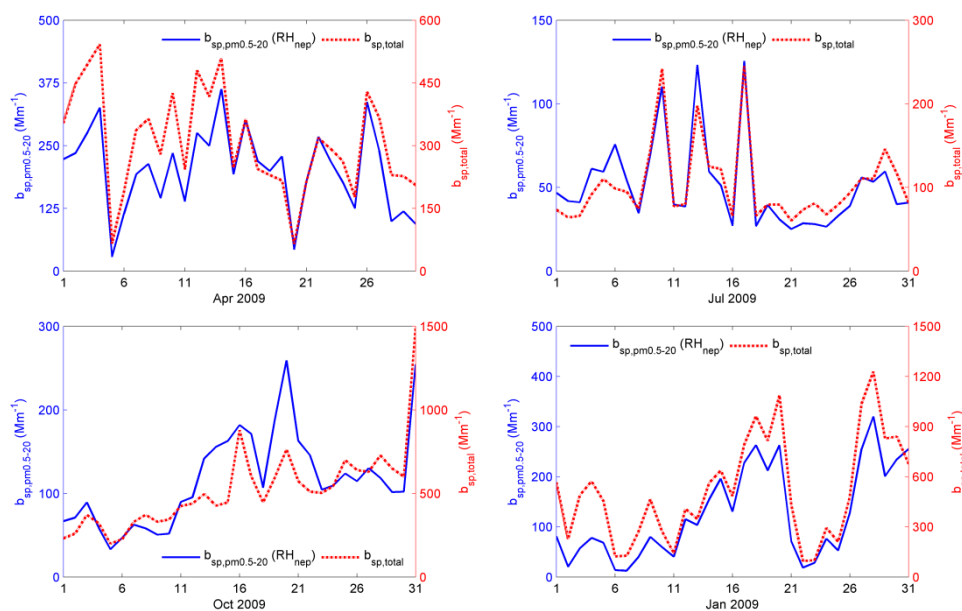


Fig. 5. Comparison between variations of estimation and measurement.

respectively. Figure 5 indicates a coherent temporal variation trend between the two data sets in April, July and January as correlation coefficient was 0.88, 0.92 and 0.94, respectively. A weaker correlation existed in October with a coefficient of 0.79. One will speculate that the practical method of calculating $b_{sp,pm0.5-20}$ in this paper can conditionally predict the variation trend of $b_{sp,total}$.

In Fig. 6, N_j of $PM_{0.5-20}$ showed its strong linear correlations with $b_{sp,j}$ and $b_{ep,j}$ of $PM_{0.5-20}$ at ambient RH condition. On the contrary, $Q_{sp,j}$ and $Q_{ep,j}$ had poor linear correlation with $b_{sp,j}$ and $b_{ep,j}$, respectively, which were determined based on the assumptions of $EORI_j$ being equaled to the $EORI_{pm2.5}$ and the $f_{eff,j}$ being equaled to the $f_{eff,pm2.5}$. Furthermore, Fig. 6 illustrates that $N_{pm0.5-20}$ had strong linear correlation with $b_{sp,pm0.5-20}$ and $b_{ep,pm0.5-20}$ as correlation coefficient was 0.92 and 0.90, respectively. In this regard, it can be inferred that those assumptions influenced $b_{sp,pm0.5-20}$ and $b_{ep,pm0.5-20}$ much less significantly than $N_{pm0.5-20}$ did, hence the practical method introduced in current study is capable of estimating variation trend of optical property of $PM_{0.5-20}$ with good accuracy, as long as the data of particles number size distribution is available and validated. Accordingly, a possible reason for the weaker correlation between $b_{sp,pm0.5-20}$ and $b_{sp,total}$ in October is that there were many even smaller particles in ambient atmosphere and APS could not distinguish them enough to produce a better particles number size distribution for the estimation. This speculation will be further investigated if a combination of SMPS-APS is introduced in future study.

4.2.2 Impact of relative humidity

In order to investigate the hygroscopicity of the optical properties of $PM_{0.5-20}$, $f_g = g(RH)/g(RH_0)$ is used to represent the hygroscopic growth factor, where $g(RH)$ stands for an optical property at a specific RH condition (denoted in fraction) and RH_0 is valued to 0.2 representing the dry state. In the light of a previous paper (Cheng et al., 2008b), a function as Eq. (8) best fitted the RH dependence curve of f_g .

$$f_g(RH) = \left(\frac{1 - RH}{1 - RH_0} \right)^{-a \cdot RH} \quad (8)$$

R^2 of the curve fittings of all samples were 0.99 on average. Temporal variations of coefficient “ a ” are plotted in Fig. 7, and judged by which, $b_{sp,pm0.5-20}$ grew more rapidly than $b_{ap,pm0.5-20}$ did as RH increased. Correlation coefficient between “ a ” of $b_{ep,pm0.5-20}$ and γ_{ion} (mass fraction of ions in $PM_{2.5}$) being 0.86 indicates that different content of ions accounted for the variation of coefficient “ a ”. Along with an increasing RH, $b_{ep,pm0.5-20}$ grew and favoured a more rapid growth when aerosol had a high content of inorganic water-soluble salts.

Regarding the average of all samples, RH dependence curves of $f_{bep,pm0.5-20}$, $f_{bsp,pm0.5-20}$ and $f_{bap,pm0.5-20}$ are also illustrated in Fig. 7 where circles stand for mean values and bars outline one time standard deviation. When RH increased from 20% to 90%, $f_{bep,pm0.5-20}$, $f_{bsp,pm0.5-20}$ and $f_{bap,pm0.5-20}$ enlarged 1.76, 1.96 and 1.19 times, respectively.

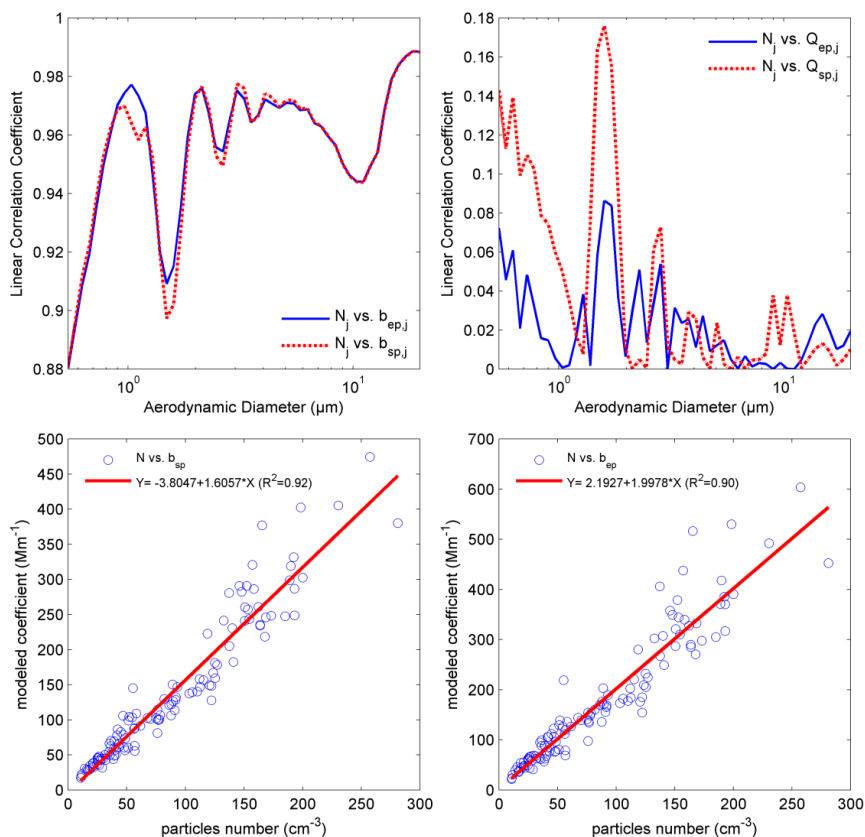


Fig. 6. Linear regressions between particles number and optical properties of $\text{PM}_{0.5-20}$.

Table 4. Potential uncertainties of the model estimation.

Property	Uncertainty
RH recorded by auto weather station ^a	3.87 %
N_j measured by APS ^b	3.30 %
$D_{a,j}$ from APS ^b	3.00 %
ion & OC/EC analysis ^c	2.49 %
$\text{PM}_{2.5}$ sample weighing ^c	1.08 %
Real part of the ORI of the EC ^b	3.00 %
Imaginary part of the ORI of the EC ^b	5.00 %
Total	8.73 %

^a From instrument technical manual,

^b noted in a previous paper (Cheng et al., 2008c),

^c estimated based on instrument detection limit.

4.2.3 Uncertainties in model estimation

Uncertainties of the model estimation discussed above are listed in Table 4, the potential sources of which includes instrumental measurement (analysis) and parameters chosen for the model. The total uncertainty was determined by Eq. (9) which was recommended by IPCC (Intergovernmental Panel on Climate Change). In Eq. (9) below, U_{total} is the total uncertainty while U_i is the i -th subset of uncertainties.

$$U_{\text{total}} = \left(\sum_i U_i^2 \right)^{1/2} \quad (9)$$

5 Summary and conclusion

In the urban area of Guangzhou, mass concentration of $\text{PM}_{2.5}$ was at its lowest level being $44.6 \mu\text{g m}^{-3}$ in July 2009, higher in April 2009 and October 2009, and reached the highest level being $130.8 \mu\text{g m}^{-3}$ in January 2010. The content of SO_4^{2-} rose in October while was low in January. The content of water uptake dropped in October for a drier climate. Moreover, NO_3^- increased its content in April and January, while POM and EC had high content in July and October.

On the basis of ions identification from $\text{PM}_{2.5}$ samples, the ISORROPIA II model helped to reconstruct aerosol chemical composition. The major species that constituted $\text{PM}_{2.5}$ particles included $(\text{NH}_4)_2\text{SO}_4$, Na_2SO_4 , K_2SO_4 , NH_4NO_3 , HNO_3 , water, POM and EC, and the monthly average contents of the major species are listed in Table 5. Regarding ambient RH, inorganic species, water, POM, EC and the Residual accounted for 34–47 %, 19–31 %, 14–20 %, 6–8 % and 8–17 % of total mass, respectively.

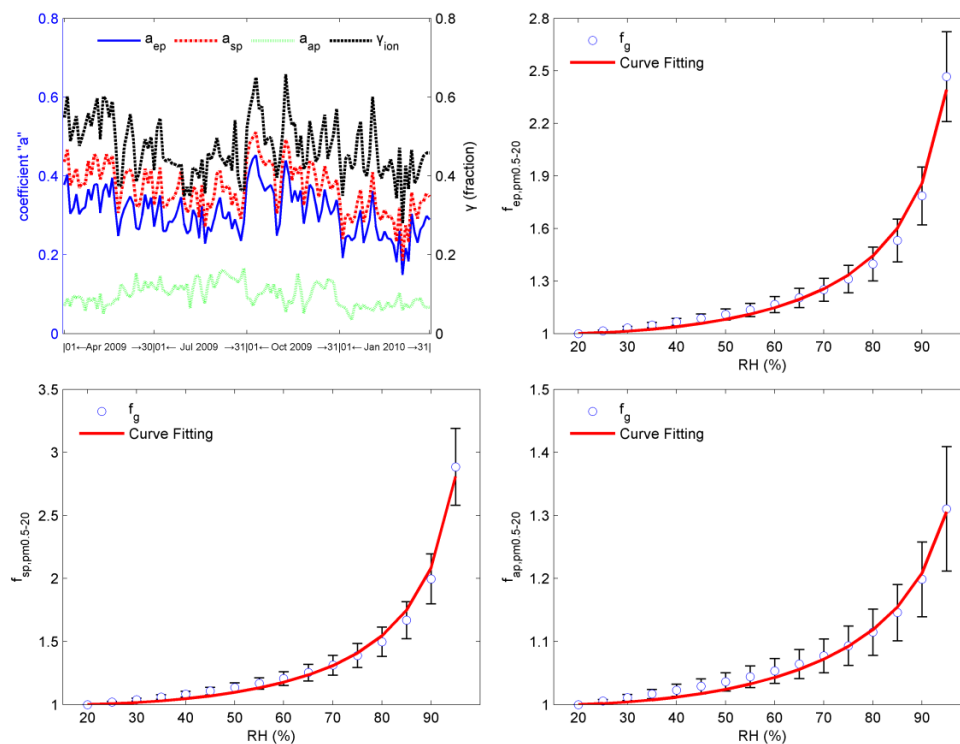


Fig. 7. Temporal variations of coefficient “*a*” and the RH dependence curves of optical properties of $\text{PM}_{0.5-20}$.

Table 5. Ambient RH condition and the corresponding chemical composition of $\text{PM}_{2.5}$.

Item	April	July	October	January
RH (in fraction)	0.68	0.68	0.50	0.72
$\text{PM}_{2.5}$ ($\mu\text{g m}^{-3}$)	94.8	44.6	95.4	130.8
Inorganic species	41 %	37 %	47 %	34 %
– $(\text{NH}_4)_2\text{SO}_4$	15.5 %	10.4 %	24.8 %	10.9 %
– Na_2SO_4	8.0 %	13.9 %	4.9 %	5.0 %
– NH_4NO_3	5.8 %	0.3 %	0.3 %	7.1 %
– K_2SO_4	2.4 %	1.4 %	2.6 %	1.8 %
– HNO_3	5.9 %	4.2 %	6.5 %	4.4 %
– NaHSO_4	0.1 %	1.2 %	3.0 %	0.0 %
– NH_4Cl	0.6 %	0.1 %	0.0 %	1.2 %
H_2O	31 %	26 %	19 %	29 %
POM	14 %	20 %	17 %	14 %
EC	6 %	8 %	7 %	6 %
Residual	8 %	10 %	9 %	17 %

Under the assumption of “internal mixture”, optical properties of $\text{PM}_{0.5-20}$ were estimated with good accuracy by the Mie Model. The EORI, EGF and ρ_{dry} were determined with an understanding of chemical composition of $\text{PM}_{2.5}$; and particles number size distribution of $\text{PM}_{0.5-20}$ was derived from APS. The monthly averages of them are summarised in Table 6.

Table 6. Ambient RH condition and the corresponding optical properties of $\text{PM}_{0.5-20}$.

Item	April	July	October	January
RH (in fraction)	0.68	0.68	0.50	0.72
n_{eff}	1.446	1.455	1.483	1.461
k_{eff}	0.034	0.041	0.040	0.032
f_{eff}	1.470	1.546	1.372	1.561
ρ_{dry} (g cm^{-3})	1.78	1.93	1.87	1.80
N (particles cm^{-3})	147.8	35.7	83.9	85.7
b_{ep} (Mm^{-1})	300.0	78.6	146.9	188.5
b_{sp} (Mm^{-1})	236.4	57.4	113.2	144.8
b_{ap} (Mm^{-1})	63.7	21.3	33.7	43.7
f_{ep}	1.29	1.23	1.14	1.26
f_{sp}	1.37	1.29	1.18	1.33
f_{ap}	1.07	1.10	1.04	1.05
ω^{a}	0.78	0.72	0.76	0.78
φ^{b}	0.66	0.49	0.23	0.23

^a ω is ratio of $b_{\text{sp},\text{pm}0.5-20}$ to $b_{\text{ep},\text{pm}0.5-20}$.

^b φ is ratio of $b_{\text{sp},\text{pm}0.5-20}$ to $b_{\text{sp},\text{total}}$.

The highest average level of $b_{\text{ep},\text{pm}0.5-20}$ being 300 Mm^{-1} happened in April while the lowest one being 78.6 Mm^{-1} in July. Regarding size distribution of $b_{\text{ep},\text{pm}0.5-20}$, peak value was almost located in the diameter range between 0.5 and $1.0 \mu\text{m}$. Furthermore, hygroscopic growth of optical

properties of $PM_{0.5-20}$ largely depended on RH. As RH increased, $b_{ep,pm0.5-20}$ grew and favoured a more rapid growth when aerosol had a high content of inorganic water-soluble salts. Averagely, $f_{bep,pm0.5-20}$ enlarged 1.76 times when RH increased from 20% to 90%. With regard to the temporal variation of ambient RH, $f_{bep,pm0.5-20}$ was 1.29, 1.23, 1.14 and 1.26 on average in April, July, October and January, respectively. Together with the measurement techniques and numerical models adopted currently, the SMPS-APS combination and the $PM_{1.0}/PM_{2.5}/PM_{10}$ sampling will be deployed in near study for a further investigation on the impact of RH and particles number size distribution on estimating aerosol optical properties.

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